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APPENDIX A. ATSDR MINIMAL RISK LEVELS AND WORKSHEETS

The Comprehensive Environmental Response, Compensation, and Liability Act (CERCLA) [42 U.S.C. 9601 et seq.], as amended by the Superfund Amendments and Reauthorization Act (SARA) [Pub. L. 99–499], requires that the Agency for Toxic Substances and Disease Registry (ATSDR) develop jointly with the U.S. Environmental Protection Agency (EPA), in order of priority, a list of hazardous substances most commonly found at facilities on the CERCLA National Priorities List (NPL); prepare toxicological profiles for each substance included on the priority list of hazardous substances; and assure the initiation of a research program to fill identified data needs associated with the substances.

The toxicological profiles include an examination, summary, and interpretation of available toxicological information and epidemiologic evaluations of a hazardous substance. During the development of toxicological profiles, Minimal Risk Levels (MRLs) are derived when reliable and sufficient data exist to identify the target organ(s) of effect or the most sensitive health effect(s) for a specific duration for a given route of exposure. An MRL is an estimate of the daily human exposure to a hazardous substance that is likely to be without appreciable risk of adverse noncancer health effects over a specified route and duration of exposure. MRLs are based on noncancer health effects only and are not based on a consideration of cancer effects. These substance-specific estimates, which are intended to serve as screening levels, are used by ATSDR health assessors to identify contaminants and potential health effects that may be of concern at hazardous waste sites. It is important to note that MRLs are not intended to define clean-up or action levels.

MRLs are derived for hazardous substances using the no-observed-adverse-effect level/uncertainty factor approach. They are below levels that might cause adverse health effects in the people most sensitive to such chemical-induced effects. MRLs are derived for acute (1–14 days), intermediate (15–364 days), and chronic (365 days and longer) durations and for the oral and inhalation routes of exposure. Currently, MRLs for the dermal route of exposure are not derived because ATSDR has not yet identified a method suitable for this route of exposure. MRLs are generally based on the most sensitive substance-induced end point considered to be of relevance to humans. Serious health effects (such as irreparable damage to the liver or kidneys, or birth defects) are not used as a basis for establishing MRLs. Exposure to a level above the MRL does not mean that adverse health effects will occur.

MRLs are intended only to serve as a screening tool to help public health professionals decide where to look more closely. They may also be viewed as a mechanism to identify those hazardous waste sites that

are not expected to cause adverse health effects. Most MRLs contain a degree of uncertainty because of the lack of precise toxicological information on the people who might be most sensitive (e.g., infants, elderly, nutritionally or immunologically compromised) to the effects of hazardous substances. ATSDR uses a conservative (i.e., protective) approach to address this uncertainty consistent with the public health principle of prevention. Although human data are preferred, MRLs often must be based on animal studies because relevant human studies are lacking. In the absence of evidence to the contrary, ATSDR assumes that humans are more sensitive to the effects of hazardous substance than animals and that certain persons may be particularly sensitive. Thus, the resulting MRL may be as much as 100-fold below levels that have been shown to be nontoxic in laboratory animals.

Proposed MRLs undergo a rigorous review process: Health Effects/MRL Workgroup reviews within the Division of Toxicology and Human Health Sciences, expert panel peer reviews, and agency-wide MRL Workgroup reviews, with participation from other federal agencies and comments from the public. They are subject to change as new information becomes available concomitant with updating the toxicological profiles. Thus, MRLs in the most recent toxicological profiles supersede previously published levels. For additional information regarding MRLs, please contact the Division of Toxicology and Human Health Sciences, Agency for Toxic Substances and Disease Registry, 1600 Clifton Road NE, Mailstop S102-1, Atlanta, Georgia 30329-4027.

MINIMAL RISK LEVEL (MRL) WORKSHEET

Chemical Name: Trichloroethylene

CAS Numbers: 79-01-6 Date: June 2019 Profile Status: Final

Route: [X] Inhalation [] Oral

Duration: [] Acute [X] Intermediate [] Chronic

Graph Key: 61, 86 Species: Mouse, Rat

Minimal Risk Level: 0.0004 [] mg/kg/day [X] ppm

The ATSDR chronic-duration inhalation MRL of 0.0004 ppm for trichloroethylene was adopted as the ATSDR intermediate-duration inhalation MRL (see chronic-duration inhalation MRL worksheet for more information). The ATSDR chronic-duration inhalation MRL is based, in part, on results of PBPK modeling exercises that simulated 100 weeks of exposure for humans (EPA 2011e). Sample simulations for a 52-week exposure (within the range of an ATSDR-defined intermediate-duration exposure [15–364 days]) resulted in the same internal dose point of departure (idPOD) as the idPOD resulting from simulations for the 100-week exposure.

Agency Contact (Chemical Manager): G. Daniel Todd, Ph.D.

MINIMAL RISK LEVEL (MRL) WORKSHEET

Chemical Name: Trichloroethylene

CAS Numbers: 79-01-6 Date: June 2019 Profile Status: Final

Route: [X] Inhalation [] Oral

Duration: [] Acute [] Intermediate [X] Chronic

Graph Key: 61, 86 Species: Mouse, Rat

Minimal Risk Level: 0.0004 [] mg/kg/day [X] ppm

The chronic-duration inhalation MRL is based on the results of two critical oral exposure studies that reported decreased thymus weight in female mice (Keil et al. 2009) and fetal heart malformations in rats (Johnson et al. 2003). In the EPA assessment for trichloroethylene (EPA 2011e), EPA developed a PBPK model, which was used to calculate the idPOD and perform route-to-route extrapolation to human equivalency concentrations (HECs) for these studies. The resulting HEC99 values were 0.033 ppm based on thymus weight and 0.0037 ppm based on fetal heart malformations. The HEC99 of 0.033 ppm for thymus weight was divided by a total uncertainty factor of 100 (to account for use of a LOAEL and to account for species extrapolation and human variability using a PBPK model); the resulting candidate chronic RfC was 0.00033 ppm. The HEC99 of 0.0037 ppm for fetal heart malformations was divided by a total uncertainty factor of 10 (to account for species extrapolation and human variability using a PBPK model); the resulting candidate chronic RfC was 0.00037 ppm. EPA (2011e) selected the midpoint value of the studies (0.0004 ppm, rounded up from 0.00035 ppm) as the chronic RfC for trichloroethylene. ATSDR agreed that this was a reasonable approach. The resulting chronic-duration inhalation MRL is 0.0004 ppm.

Keil et al. (2009)

<u>Reference</u>: Keil DE, Peden-Adams MM, Wallace S, et al. 2009. Assessment of trichloroethylene (TCE) exposure in murine strains genetically-prone and non-prone to develop autoimmune disease. J Environ Sci Health A Tox Hazard Subst Environ Eng 44(5):443-453.

Experimental design: Groups of 9-week-old female B6C3F1 mice (9–10/group) were administered trichloroethylene in the drinking water at 0, 1,400, or 14,000 ppb (1.4 or 14 ppm) in 1% emulphor vehicle for 30 weeks. During the exposure period, serum levels of total IgG and autoantibodies (anti-ssDNA, -dsDNA, and -glomerular antigen [GA]) were monitored. Body weights were recorded 1 day prior to the initiation of trichloroethylene exposure and again at exposure termination. At sacrifice, the spleen, thymus, liver, and kidneys were weighed. Spleen and thymus were processed for assessment of cell counts and activity. Kidneys were processed for histopathologic evaluation; renal pathology was scored by grading glomerular inflammation, crescent formation, and necrosis in histopathology slides.

<u>Effect noted in study and corresponding doses</u>: Decreased thymus weight (30% lower than controls) and increased serum levels of IgG and selected autoantibodies at 1.4 ppm trichloroethylene in the drinking water (EPA-estimated dose of 0.35 mg/kg/day).

<u>Dose and end point used for MRL derivation</u>: A PBPK model was used to calculate the idPOD (idPOD=0.139 mg trichloroethylene metabolized/kg $^{3/4}$ /day) from the applied dose LOAEL of 0.35 mg/kg/day. The mouse idPOD was converted to a HEC $_{99}$ (HEC $_{99}$ =0.033 ppm) for lifetime

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continuous exposure derived from combined interspecies, intraspecies, and route-to-route extrapolation using the PBPK model for trichloroethylene.

[] NOAEL [] LOAEL [X] HEC99

Uncertainty Factors used in MRL derivation:

- [X] 10 for use of a LOAEL
- [X] 3.16 for extrapolation from animals to humans because a PBPK model was used
- [X] 3.16 for human variability because a PBPK model was used to characterize human toxicokinetic variability

Was a conversion factor used from ppm in food or water to a mg/body weight dose? EPA estimated doses using the average of subchronic and chronic reference values for generic body weight and water consumption rates for female B6C3F1 mice.

If an inhalation study in animals, list conversion factors used in determining human equivalent dose: Not applicable.

Was a conversion used from intermittent to continuous exposure? Not applicable.

Johnson et al. (2003)

<u>Reference</u>: Johnson PD, Goldberg SY, Mays MZ, et al. 2003. Threshold of trichloroethylene contamination in maternal drinking waters affecting fetal heart development in the rat. Environ Health Perspect 111(3):289-292.

Experimental design: Groups of pregnant Sprague-Dawley rats (9–13/exposure level) were administered trichloroethylene in the drinking water throughout gestation (GDs 1–22) at concentrations of 0, 0.0025, 0.25, 1.5, or 1,000 ppm. At termination on GD 22, dams and fetuses were examined for gross abnormalities and fetuses were weighed, measured for crown-rump length, and sexed. Fetal hearts and great vessels were examined for gross malformations and prepared for histopathologic evaluations.

Effect noted in study and corresponding doses: Increased incidences of fetuses with cardiac malformations at maternal exposure levels \geq 0.25 ppm (estimated maternal doses \geq 0.048 mg/kg/day).

Dose and end point used for MRL derivation: Using a benchmark response (BMR) of 1% extra risk that was preferred due to accounting for intralitter effects using a nested model, and pups being the unit of measure, EPA (2011e) calculated a rat lower 95% confidence limit on the benchmark dose (BMDL₀₁) of 0.0207 mg/kg/day from the fetal heart malformation incidence data. The highest dose group (1,000-fold higher than next highest) was dropped to improve model fit. The rat BMDL₀₁ was 0.0207 mg/kg/day. A PBPK model was used to calculate the idPOD of 0.0142 mg trichloroethylene metabolized by oxidation/kg body weight^{3/4}/day. The rat idPOD was converted to a HEC₉₉ of 0.0037 ppm for continuous lifetime exposure derived from route-to-route extrapolation and combined interspecies and intraspecies extrapolation using the PBPK model.

[] NOAEL [] LOAEL [X] HEC99

Uncertainty Factors used in MRL derivation:

[X] 3.16 for extrapolation from animals to humans because a PBPK model was used

[X] 3.16 for human variability because a PBPK model was used to characterize human toxicokinetic variability

Was a conversion factor used from ppm in food or water to a mg/body weight dose? No.

If an inhalation study in animals, list conversion factors used in determining human equivalent dose: Not applicable.

Was a conversion used from intermittent to continuous exposure? Not applicable.

Other additional studies or pertinent information that lend support to this MRL: Studies in humans and animals have identified a number of potential targets of trichloroethylene toxicity, including the central nervous system, liver, kidney, immune system, male reproductive system, and the developing fetus. The toxicity of trichloroethylene does not appear to be route-specific; similar effects have been observed following inhalation and oral exposure. The most sensitive toxicity targets appear to be the immune system, the developing fetus, and the kidneys. Immunological effects include alterations in thymus weight and immune function. Decreases in thymus weights were observed in mice exposed to 0.35 mg/kg/day trichloroethylene in drinking water for 30 weeks (Keil et al. 2009) and an impaired response to SRBCs and delayed hypersensitivity were observed in the pups of mice exposed to 0.37 mg/kg/day trichloroethylene in drinking water on GDs 0-21 and for 3 or 8 weeks postpartum (Peden-Adams et al. 2006). At higher doses, other studies have reported increases in splenic lymphocytes (Blossom and Doss 2007) and impaired immune function (Blossom et al. 2008; Sanders et al. 1982). Developmental toxicity studies have demonstrated a number of effects including increases in the occurrence of cardiac malformations in the offspring of rats exposed to 0.048 mg/kg/day trichloroethylene on GDs 0-21 (Johnson et al. 2003) or 0.218 mg/kg/day trichloroethylene in drinking water prior to mating and during gestation (Dawson et al. 1993; Johnson et al. 1998); at higher doses, neurobehavioral effects, decreases in pup body weight, and perinatal mortality have been observed (Manson et al. 1984; NTP 1986; Taylor et al. 1985). Observed renal effects include toxic nephropathy in rats administered via gavage 500 mg/kg trichloroethylene 5 days/week for 2 years (NTP 1988).

EPA employed a PBPK model to calculate an idPOD for plausible internal dose-metrics based on present understanding of the role different trichloroethylene metabolites play in trichloroethylene toxicity and the mode of action for toxicity. The PBPK model was used to estimate interspecies and intraspecies pharmacokinetic variability and resulted in HEC₉₉ values for candidate critical effects.

Agency Contact (Chemical Manager): G. Daniel Todd, Ph.D.

MINIMAL RISK LEVEL (MRL) WORKSHEET

Chemical Name: Trichloroethylene

CAS Numbers: 79-01-6 Date: June 2019 Profile Status: Final

Route: [] Inhalation [X] Oral

Duration: [] Acute [X] Intermediate [] Chronic

Graph Key: 54, 56, 69 Species: Mouse, Rat

Minimal Risk Level: 0.0005 [X] mg/kg/day [] ppm

The ATSDR chronic-duration oral MRL of 0.0005 mg/kg/day for trichloroethylene was adopted as the ATSDR intermediate-duration oral MRL (see chronic-duration oral MRL worksheet for more information). The ATSDR chronic-duration oral MRL is based, in part, on results of PBPK modeling exercises that simulated 100 weeks of exposure for humans (EPA 2011e). Sample simulations for a 52-week exposure (within the range of an ATSDR-defined intermediate-duration exposure [15–364 days]) resulted in the same idPOD as the idPOD resulting from simulations for the 100-week exposure.

Agency Contact (Chemical Manager): G. Daniel Todd, Ph.D.

MINIMAL RISK LEVEL (MRL) WORKSHEET

Chemical Name: Trichloroethylene

CAS Numbers: 79-01-6 Date: June 2019 Profile Status: Final

Route: [] Inhalation [X] Oral

Duration: [] Acute [] Intermediate [X] Chronic

Graph Key: 54, 56, 69 Species: Mouse, Rat

Minimal Risk Level: 0.0005 [X] mg/kg/day [] ppm

The chronic-duration oral MRL is based on the results of three critical oral exposure studies that reported immunotoxicity (decreased plaque-forming cell response and increased delayed-type hypersensitivity) in mice (Peden-Adams et al. 2006), decreased thymus weight in female mice (Keil et al. 2009), and fetal heart malformations in rats (Johnson et al. 2003). In the EPA assessment for trichloroethylene (EPA 2011e), independent candidate chronic RfD values were calculated for each of these effects. The Peden-Adams et al. (2006) immunotoxicity LOAEL of 0.37 mg/kg/day was divided by a total uncertainty factor of 1,000 (to account for use of a LOAEL, interspecies extrapolation, and human variability), resulting in a candidate chronic RfD of 0.00037 mg/kg/day. The Keil et al. (2009) thymus weight LOAEL of 0.35 mg/kg/day was used to derive a PBPK model-based human equivalent dose (HED99) of 0.048 mg/kg/day, which was divided by a total uncertainty factor of 100 (to account for use of a LOAEL, interspecies extrapolation, and human variability using a PBPK model), resulting in a candidate chronic RfD of 0.00048 mg/kg/day. The Johnson et al. (2003) fetal heart malformation data were subjected to benchmark dose analysis. The resulting BMDL₀₁ (1% extra risk) of 0.0207 mg/kg/day was used to calculate a PBPK model-based HED99 of 0.0051 mg/kg/day, which was divided by a total uncertainty factor of 10 (to account for interspecies extrapolation and human variability using a PBPK model). The resulting candidate chronic RfD was 0.00051 mg/kg/day. EPA (2011e) elected to use a chronic RfD value of 0.0005 mg/kg/day and noted that this value was supported by results for multiple effects. ATSDR agreed that this was a reasonable approach. Therefore, the chronic-duration oral MRL is 0.0005 mg/kg/day.

Peden-Adams et al. (2006)

<u>Reference</u>: Peden-Adams MM, Eudaly JG, Heesemann LM, et al. 2006. Developmental immunotoxicity of trichloroethylene (TCE): studies in B6C3F1 mice. J Environ Sci Health A Tox Hazard Subst Environ Eng 41(3):249-271.

Experimental design: Groups of C3H/HeJ male and C57BL/6N female mice (5/sex/group) were administered trichloroethylene in the drinking water at 0, 1.4, or 14 ppm, beginning at pairing (1:1) and continuing for 7 days of mating and throughout gestation (at least for the dams) and lactation. Pups (strain B6C3F1 is produced from the paired parental strains) were evaluated for body length (crown-rump), and timing of eye opening and ear unfolding. At weaning of the pups at 3 weeks of age, 5—7 pups/treatment group, were weighed and sacrificed to assess kidney, liver, thymus, and spleen weights. Trichloroethylene-related effects on the immune system were assessed by measuring splenic lymphocyte proliferation, NK cell activity, SRBC-specific IgM production (PFC response), splenic B220+ cells, and thymic and splenic T-cell immunophenotypes. The remaining pups (4–5 pups/treatment group) were assessed at 8 weeks of age in a manner similar to those assessed at 3 weeks of age, with additional assessments of autoantibodies to dsDNA and delayed type hypersensitivity response (indicated by foot pad swelling following subcutaneous injection of SRBC).

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Effect noted in study and corresponding doses: Decreased PFC response was observed in 3- and 8-week-old pups and increased delayed-type sensitivity was noted in 8-week-old pups at 1.4 and 14 ppm trichloroethylene in the drinking water (author-estimated maternal doses of 0.37 and 3.7 mg/kg/day, respectively).

<u>Dose and end point used for MRL derivation</u>: 0.37 mg/kg/day for decreased PFC response and increased delayed-type sensitivity.

[] NOAEL [X] LOAEL

Uncertainty Factors used in MRL derivation:

- [X] 10 for use of a LOAEL
- [X] 10 to account for extrapolation from animals to humans
- [X] 10 to account for human variability

Was a conversion factor used from ppm in food or water to a mg/body weight dose? Doses were estimated by the study authors.

If an inhalation study in animals, list conversion factors used in determining human equivalent dose: Not applicable.

Was a conversion used from intermittent to continuous exposure? Not applicable.

Keil et al. (2009)

<u>Reference</u>: Keil DE, Peden-Adams MM, Wallace S, et al. 2009. Assessment of trichloroethylene (TCE) exposure in murine strains genetically-prone and non-prone to develop autoimmune disease. J Environ Sci Health A Tox Hazard Subst Environ Eng 44(5):443-453.

Experimental design: Groups of 9-week-old female B6C3F1 mice (9–10/group) were administered trichloroethylene in the drinking water at 0, 1,400, or 14,000 ppb (1.4 or 14 ppm) in 1% emulphor vehicle for 30 weeks. During the exposure period, serum levels of total IgG and autoantibodies (anti-ssDNA, -dsDNA, and -GA) were monitored. Body weights were recorded 1 day prior to the initiation of trichloroethylene exposure and again at exposure termination. At sacrifice, the spleen, thymus, liver, and kidneys were weighed. Spleen and thymus were processed for assessment of cell counts and activity. Kidneys were processed for histopathologic evaluation; renal pathology was scored by grading glomerular inflammation, crescent formation, and necrosis in histopathology slides.

<u>Effect noted in study and corresponding doses</u>: Decreased thymus weight (30% lower than controls) and increased serum levels of IgG and selected autoantibodies at 1.4 ppm trichloroethylene in the drinking water (EPA-estimated dose of 0.35 mg/kg/day).

Dose and end point used for MRL derivation: A PBPK model was used to calculate the idPOD (idPOD=0.139 mg trichloroethylene metabolized/kg³/4/day) from the applied dose LOAEL of 0.35 mg/kg/day. The mouse idPOD was converted to a HED99 of 0.048 mg/kg/day for lifetime continuous exposure derived from combined interspecies and intraspecies extrapolation using the PBPK model for trichloroethylene.

[] NOAEL [] LOAEL [X] HED99

Uncertainty Factors used in MRL derivation:

- [X] 10 for use of a LOAEL
- [X] 3.16 for extrapolation from animals to humans because a PBPK model was used
- [X] 3.16 for human variability because a PBPK model was used to characterize human toxicokinetic variability

Was a conversion factor used from ppm in food or water to a mg/body weight dose? EPA estimated doses using the average of subchronic and chronic reference values for generic body weight and water consumption rates for female B6C3F1 mice.

If an inhalation study in animals, list conversion factors used in determining human equivalent dose: Not applicable.

Was a conversion used from intermittent to continuous exposure? Not applicable.

Johnson et al. (2003)

<u>Reference</u>: Johnson PD, Goldberg SY, Mays MZ, et al. 2003. Threshold of trichloroethylene contamination in maternal drinking waters affecting fetal heart development in the rat. Environ Health Perspect 111(3):289-292.

Experimental design: Groups of pregnant Sprague-Dawley rats (9–13/exposure level) were administered trichloroethylene in the drinking water throughout gestation (GDs 1–22) at concentrations of 0, 0.0025, 0.25, 1.5, or 1,000 ppm. At termination on GD 22, dams and fetuses were examined for gross abnormalities and fetuses were weighed, measured for crown-rump length, and sexed. Fetal hearts and great vessels were examined for gross malformations and prepared for histopathologic evaluations.

Effect noted in study and corresponding doses: Increased incidences of fetuses with cardiac malformations at maternal exposure levels \geq 0.25 ppm (estimated maternal doses \geq 0.048 mg/kg/day).

Dose and end point used for MRL derivation: Using a BMR of 1% extra risk that was preferred due to accounting for intralitter effects using a nested model, and pups being the unit of measure, EPA (2011e) calculated a rat BMDL $_{01}$ of 0.0207 mg/kg/day from the fetal heart malformation incidence data. The highest dose group (1,000-fold higher than next highest) was dropped to improve model fit. The rat BMDL $_{01}$ was 0.0207 mg/kg/day. A PBPK model was used to calculate the idPOD of 0.0142 mg trichloroethylene metabolized by oxidation/kg body weight^{3/4}/day. The rat idPOD was converted to a HED $_{99}$ of 0.0051 mg/kg/day for continuous lifetime exposure derived from combined interspecies and intraspecies extrapolation using the PBPK model for trichloroethylene.

[] NOAEL [] LOAEL [X] HED₉₉

<u>Uncertainty Factors used in MRL derivation</u>:

- [X] 3.16 for extrapolation from animals to humans because a PBPK model was used
- [X] 3.16 for human variability because a PBPK model was used to characterize human toxicokinetic variation

Was a conversion factor used from ppm in food or water to a mg/body weight dose? No.

If an inhalation study in animals, list conversion factors used in determining human equivalent dose: Not applicable.

Was a conversion used from intermittent to continuous exposure? Not applicable.

Other additional studies or pertinent information that lend support to this MRL: Studies in humans and animals have identified a number of potential targets of trichloroethylene toxicity including the central nervous system, liver, kidney, immune system, male reproductive system, and the developing fetus. The toxicity of trichloroethylene does not appear to be route-specific; similar effects have been observed following inhalation and oral exposure. The most sensitive toxicity targets appear to be the immune system, the developing fetus, and the kidneys. Immunological effects include alterations in thymus weight and immune function. Decreases in thymus weights were observed in mice exposed to 0.35 mg/kg/day trichloroethylene in drinking water for 30 weeks (Keil et al. 2009) and an impaired response to SRBCs and delayed hypersensitivity were observed in the pups of mice exposed to 0.37 mg/kg/day trichloroethylene in drinking water on GDs 0-21 and for 3 or 8 weeks postpartum (Peden-Adams et al. 2006). At higher doses, other studies have reported increases in splenic lymphocytes (Blossom and Doss 2007) and impaired immune function (Blossom et al. 2008; Sanders et al. 1982). Developmental toxicity studies have demonstrated a number of effects including increases in the occurrence of cardiac malformations in the offspring of rats exposed to 0.048 mg/kg/day trichloroethylene on GDs 0-21 (Johnson et al. 2003) or 0.218 mg/kg/day trichloroethylene in drinking water prior to mating and during gestation (Dawson et al. 1993; Johnson et al. 1998); at higher doses, neurobehavioral effects, decreases in pup body weight, and perinatal mortality have been observed (Manson et al. 1984; NTP 1986; Taylor et al. 1985). Observed renal effects include toxic nephropathy in rats administered via gavage 500 mg/kg trichloroethylene 5 days/week for 2 years (NTP 1988).

EPA employed a PBPK model to calculate an idPOD for plausible internal dose-metrics based on present understanding of the role different trichloroethylene metabolites play in trichloroethylene toxicity and the mode of action for toxicity. The PBPK model was used to estimate interspecies and intraspecies pharmacokinetic variability and resulted in HED₉₉ values for candidate critical effects.

Agency Contact (Chemical Manager): G. Daniel Todd, Ph.D.

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APPENDIX B. USER'S GUIDE

Chapter 1

Public Health Statement

This chapter of the profile is a health effects summary written in non-technical language. Its intended audience is the general public, especially people living in the vicinity of a hazardous waste site or chemical release. If the Public Health Statement were removed from the rest of the document, it would still communicate to the lay public essential information about the chemical.

The major headings in the Public Health Statement are useful to find specific topics of concern. The topics are written in a question and answer format. The answer to each question includes a sentence that will direct the reader to chapters in the profile that will provide more information on the given topic.

Chapter 2

Relevance to Public Health

This chapter provides a health effects summary based on evaluations of existing toxicologic, epidemiologic, and toxicokinetic information. This summary is designed to present interpretive, weight-of-evidence discussions for human health end points by addressing the following questions:

- 1. What effects are known to occur in humans?
- 2. What effects observed in animals are likely to be of concern to humans?
- 3. What exposure conditions are likely to be of concern to humans, especially around hazardous waste sites?

The chapter covers end points in the same order that they appear within the Discussion of Health Effects by Route of Exposure section, by route (inhalation, oral, and dermal) and within route by effect. Human data are presented first, then animal data. Both are organized by duration (acute, intermediate, chronic). *In vitro* data and data from parenteral routes (intramuscular, intravenous, subcutaneous, etc.) are also considered in this chapter.

The carcinogenic potential of the profiled substance is qualitatively evaluated, when appropriate, using existing toxicokinetic, genotoxic, and carcinogenic data. ATSDR does not currently assess cancer potency or perform cancer risk assessments. Minimal Risk Levels (MRLs) for noncancer end points (if derived) and the end points from which they were derived are indicated and discussed.

Limitations to existing scientific literature that prevent a satisfactory evaluation of the relevance to public health are identified in the Chapter 3 Data Needs section.

Interpretation of Minimal Risk Levels

Where sufficient toxicologic information is available, ATSDR has derived MRLs for inhalation and oral routes of entry at each duration of exposure (acute, intermediate, and chronic). These MRLs are not meant to support regulatory action, but to acquaint health professionals with exposure levels at which adverse health effects are not expected to occur in humans.

MRLs should help physicians and public health officials determine the safety of a community living near a hazardous substance emission, given the concentration of a contaminant in air or the estimated daily dose in water. MRLs are based largely on toxicological studies in animals and on reports of human occupational exposure.

MRL users should be familiar with the toxicologic information on which the number is based. Chapter 2, "Relevance to Public Health," contains basic information known about the substance. Other sections such as Chapter 3 Section 3.9, "Interactions with Other Substances," and Section 3.10, "Populations that are Unusually Susceptible" provide important supplemental information.

MRL users should also understand the MRL derivation methodology. MRLs are derived using a modified version of the risk assessment methodology that the Environmental Protection Agency (EPA) provides (Barnes and Dourson 1988) to determine reference doses (RfDs) for lifetime exposure.

To derive an MRL, ATSDR generally selects the most sensitive end point which, in its best judgement, represents the most sensitive human health effect for a given exposure route and duration. ATSDR cannot make this judgement or derive an MRL unless information (quantitative or qualitative) is available for all potential systemic, neurological, and developmental effects. If this information and reliable quantitative data on the chosen end point are available, ATSDR derives an MRL using the most sensitive species (when information from multiple species is available) with the highest no-observed-adverse-effect level (NOAEL) that does not exceed any adverse effect levels. When a NOAEL is not available, a lowest-observed-adverse-effect level (LOAEL) can be used to derive an MRL, and an uncertainty factor (UF) of 10 must be employed. Additional uncertainty factors of 10 must be used both for human variability to protect sensitive subpopulations (people who are most susceptible to the health effects caused by the substance) and for interspecies variability (extrapolation from animals to humans). In deriving an MRL, these individual uncertainty factors are multiplied together. The product is then divided into the inhalation concentration or oral dosage selected from the study. Uncertainty factors used in developing a substance-specific MRL are provided in the footnotes of the levels of significant exposure (LSE) tables.

Chapter 3

Health Effects

Tables and Figures for Levels of Significant Exposure (LSE)

Tables and figures are used to summarize health effects and illustrate graphically levels of exposure associated with those effects. These levels cover health effects observed at increasing dose concentrations and durations, differences in response by species, MRLs to humans for noncancer end points, and EPA's estimated range associated with an upper- bound individual lifetime cancer risk of 1 in 10,000 to 1 in 10,000,000. Use the LSE tables and figures for a quick review of the health effects and to locate data for a specific exposure scenario. The LSE tables and figures should always be used in conjunction with the text. All entries in these tables and figures represent studies that provide reliable, quantitative estimates of NOAELs, LOAELs, or Cancer Effect Levels (CELs).

The legends presented below demonstrate the application of these tables and figures. Representative examples of LSE Table 3-1 and Figure 3-1 are shown. The numbers in the left column of the legends correspond to the numbers in the example table and figure.

LEGEND

See Sample LSE Table 3-1 (page B-6)

- (1) Route of Exposure. One of the first considerations when reviewing the toxicity of a substance using these tables and figures should be the relevant and appropriate route of exposure. Typically when sufficient data exist, three LSE tables and two LSE figures are presented in the document. The three LSE tables present data on the three principal routes of exposure, i.e., inhalation, oral, and dermal (LSE Tables 3-1, 3-2, and 3-3, respectively). LSE figures are limited to the inhalation (LSE Figure 3-1) and oral (LSE Figure 3-2) routes. Not all substances will have data on each route of exposure and will not, therefore, have all five of the tables and figures.
- (2) <u>Exposure Period</u>. Three exposure periods—acute (less than 15 days), intermediate (15–364 days), and chronic (365 days or more)—are presented within each relevant route of exposure. In this example, an inhalation study of intermediate exposure duration is reported. For quick reference to health effects occurring from a known length of exposure, locate the applicable exposure period within the LSE table and figure.
- (3) <u>Health Effect</u>. The major categories of health effects included in LSE tables and figures include death, systemic, immunological, neurological, developmental, reproductive, and cancer.

 NOAELs and LOAELs can be reported in the tables and figures for all effects but cancer.

 Systemic effects are further defined in the "System" column of the LSE table (see key number 18).
- (4) <u>Key to Figure</u>. Each key number in the LSE table links study information to one or more data points using the same key number in the corresponding LSE figure. In this example, the study represented by key number 18 has been used to derive a NOAEL and a Less Serious LOAEL (also see the two "18r" data points in sample Figure 3-1).
- (5) Species. The test species, whether animal or human, are identified in this column. Chapter 2, "Relevance to Public Health," covers the relevance of animal data to human toxicity and Section 3.4, "Toxicokinetics," contains any available information on comparative toxicokinetics. Although NOAELs and LOAELs are species specific, the levels are extrapolated to equivalent human doses to derive an MRL.
- (6) Exposure Frequency/Duration. The duration of the study and the weekly and daily exposure regimens are provided in this column. This permits comparison of NOAELs and LOAELs from different studies. In this case (key number 18), rats were exposed to "Chemical x" via inhalation for 6 hours/day, 5 days/week, for 13 weeks. For a more complete review of the dosing regimen, refer to the appropriate sections of the text or the original reference paper (i.e., Nitschke et al. 1981).
- (7) System. This column further defines the systemic effects. These systems include respiratory, cardiovascular, gastrointestinal, hematological, musculoskeletal, hepatic, renal, and dermal/ocular. "Other" refers to any systemic effect (e.g., a decrease in body weight) not covered in these systems. In the example of key number 18, one systemic effect (respiratory) was investigated.
- (8) <u>NOAEL</u>. A NOAEL is the highest exposure level at which no adverse effects were seen in the organ system studied. Key number 18 reports a NOAEL of 3 ppm for the respiratory system, which was used to derive an intermediate exposure, inhalation MRL of 0.005 ppm (see footnote "b").

- (9) <u>LOAEL</u>. A LOAEL is the lowest dose used in the study that caused an adverse health effect. LOAELs have been classified into "Less Serious" and "Serious" effects. These distinctions help readers identify the levels of exposure at which adverse health effects first appear and the gradation of effects with increasing dose. A brief description of the specific end point used to quantify the adverse effect accompanies the LOAEL. The respiratory effect reported in key number 18 (hyperplasia) is a Less Serious LOAEL of 10 ppm. MRLs are not derived from Serious LOAELs.
- (10) Reference. The complete reference citation is given in Chapter 9 of the profile.
- (11) <u>CEL</u>. A CEL is the lowest exposure level associated with the onset of carcinogenesis in experimental or epidemiologic studies. CELs are always considered serious effects. The LSE tables and figures do not contain NOAELs for cancer, but the text may report doses not causing measurable cancer increases.
- (12) <u>Footnotes</u>. Explanations of abbreviations or reference notes for data in the LSE tables are found in the footnotes. Footnote "b" indicates that the NOAEL of 3 ppm in key number 18 was used to derive an MRL of 0.005 ppm.

LEGEND

See Sample Figure 3-1 (page B-7)

LSE figures graphically illustrate the data presented in the corresponding LSE tables. Figures help the reader quickly compare health effects according to exposure concentrations for particular exposure periods.

- (13) <u>Exposure Period</u>. The same exposure periods appear as in the LSE table. In this example, health effects observed within the acute and intermediate exposure periods are illustrated.
- (14) <u>Health Effect</u>. These are the categories of health effects for which reliable quantitative data exists. The same health effects appear in the LSE table.
- (15) <u>Levels of Exposure</u>. Concentrations or doses for each health effect in the LSE tables are graphically displayed in the LSE figures. Exposure concentration or dose is measured on the log scale "y" axis. Inhalation exposure is reported in mg/m³ or ppm and oral exposure is reported in mg/kg/day.
- (16) <u>NOAEL</u>. In this example, the open circle designated 18r identifies a NOAEL critical end point in the rat upon which an intermediate inhalation exposure MRL is based. The key number 18 corresponds to the entry in the LSE table. The dashed descending arrow indicates the extrapolation from the exposure level of 3 ppm (see entry 18 in the table) to the MRL of 0.005 ppm (see footnote "b" in the LSE table).
- (17) <u>CEL</u>. Key number 38m is one of three studies for which CELs were derived. The diamond symbol refers to a CEL for the test species-mouse. The number 38 corresponds to the entry in the LSE table.

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- (18) <u>Estimated Upper-Bound Human Cancer Risk Levels</u>. This is the range associated with the upper-bound for lifetime cancer risk of 1 in 10,000 to 1 in 10,000,000. These risk levels are derived from the EPA's Human Health Assessment Group's upper-bound estimates of the slope of the cancer dose response curve at low dose levels (q₁*).
- (19) <u>Key to LSE Figure</u>. The Key explains the abbreviations and symbols used in the figure.

SAMPLE

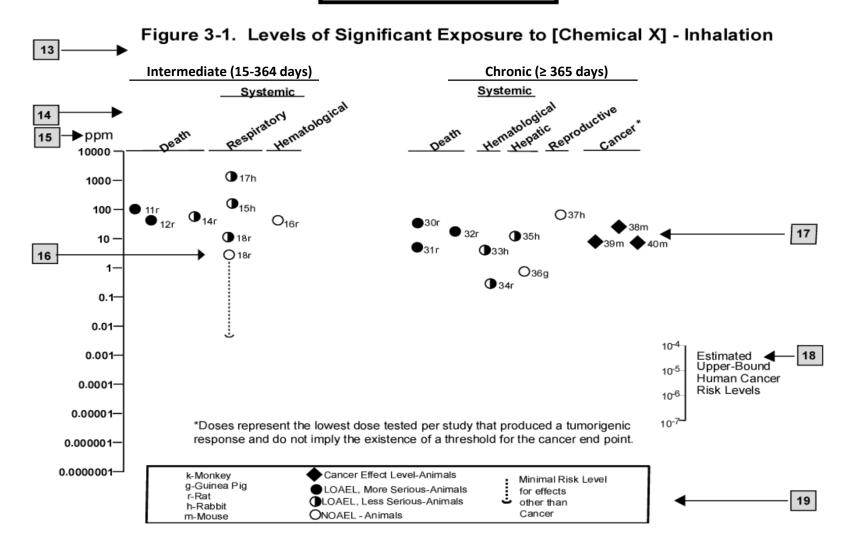
Table 3-1. Levels of Significant Exposure to [Chemical x] – Inhalation

| | | Key to figure ^a | Species | Exposure frequency/s duration | System | NOAEL (ppm) | LOAEL (eff Less seriou (ppm) | | Serious (ppm) | Reference |
|---|---------------|----------------------------|--------------|-------------------------------|----------|----------------|------------------------------------|--------------|--------------------------------------|------------------|
| 2 | \rightarrow | INTERMEDIATE EXPOSURE | | | | | | | _ | |
| | | | 5 | 6 | 7 | 8 | 9 | | | 10 |
| 3 | \rightarrow | Systemic | \downarrow | ↓ | \ | \downarrow | \downarrow | | | \ |
| 4 | \rightarrow | 18 | Rat | 13 wk 5 d/wk 6 hr/d | Resp | 3 ^b | 10 (hyperplasia) | | Nitschke et al. 1981 | |
| П | | CHRONIC EXPOSURE | | | | | | | | |
| | | Cancer | | | | | | 11 | | |
| | | | | | | | | \downarrow | | |
| | | 38 | Rat | 18 mo 5 d/wk 7 hr/d | | | | 20 | (CEL, multiple organs) | Wong et al. 1982 |
| | | 39 | Rat | 89–104 wk 5 d/wk 6 hr/d | | | | 10 | (CEL, lung tumors, nasal tumors) | NTP 1982 |
| | | 40 | Mouse | 79–103 wk 5 d/wk 6 hr/d | | | | 10 | (CEL, lung tumors, hemangiosarcomas) | NTP 1982 |

^{2 →} a The number corresponds to entries in Figure 3-1.

^b Used to derive an intermediate inhalation Minimal Risk Level (MRL) of 5x10⁻³ ppm; dose adjusted for intermittent exposure and divided by an uncertainty factor of 100 (10 for extrapolation from animal to humans, 10 for human variability).

SAMPLE



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APPENDIX C. ACRONYMS, ABBREVIATIONS, AND SYMBOLS

ACGIH American Conference of Governmental Industrial Hygienists
ACOEM American College of Occupational and Environmental Medicine

ADI acceptable daily intake

ADME absorption, distribution, metabolism, and excretion

AED atomic emission detection
AFID alkali flame ionization detector
AFOSH Air Force Office of Safety and Health

ALT alanine aminotransferase AML acute myeloid leukemia

AOAC Association of Official Analytical Chemists

AOEC Association of Occupational and Environmental Clinics

AP alkaline phosphatase

APHA American Public Health Association

AST aspartate aminotransferase

atm atmosphere

ATSDR Agency for Toxic Substances and Disease Registry

AWQC Ambient Water Quality Criteria
BAT best available technology
BCF bioconcentration factor
BEI Biological Exposure Index

BMD/C benchmark dose or benchmark concentration

BMD_X dose that produces a X% change in response rate of an adverse effect

BMDL_X 95% lower confidence limit on the BMD_X

BMDS Benchmark Dose Software BMR benchmark response

BSC Board of Scientific Counselors

C centigrade CAA Clean Air Act

CAG Cancer Assessment Group of the U.S. Environmental Protection Agency

CAS Chemical Abstract Services

CDC Centers for Disease Control and Prevention

CEL cancer effect level

CELDS Computer-Environmental Legislative Data System

CERCLA Comprehensive Environmental Response, Compensation, and Liability Act

CFR Code of Federal Regulations

Ci curie

CI confidence interval

CLP Contract Laboratory Program

cm centimeter

CML chronic myeloid leukemia

CPSC Consumer Products Safety Commission

CWA Clean Water Act

DHEW Department of Health, Education, and Welfare DHHS Department of Health and Human Services

DNA deoxyribonucleic acid DOD Department of Defense DOE Department of Energy DOL Department of Labor

DOT Department of Transportation

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DOT/UN/ Department of Transportation/United Nations/

NA/IMDG North America/Intergovernmental Maritime Dangerous Goods Code

DWEL drinking water exposure level ECD electron capture detection

ECG/EKG electrocardiogram
EEG electroencephalogram

EEGL Emergency Exposure Guidance Level EPA Environmental Protection Agency

F Fahrenheit

F₁ first-filial generation

FAO Food and Agricultural Organization of the United Nations

FDA Food and Drug Administration

FEMA Federal Emergency Management Agency

FIFRA Federal Insecticide, Fungicide, and Rodenticide Act

FPD flame photometric detection

fpm feet per minute FR Federal Register

FSH follicle stimulating hormone

g gram

GC gas chromatography gd gestational day

GLC gas liquid chromatography GPC gel permeation chromatography

HPLC high-performance liquid chromatography
HRGC high resolution gas chromatography
HSDB Hazardous Substance Data Bank

IARC International Agency for Research on Cancer IDLH immediately dangerous to life and health

ILO International Labor Organization
IRIS Integrated Risk Information System

Kd adsorption ratio kg kilogram

kkg kilokilogram; 1 kilokilogram is equivalent to 1,000 kilograms and 1 metric ton

 K_{oc} organic carbon partition coefficient K_{ow} octanol-water partition coefficient

L liter

 $\begin{array}{lll} LC & liquid chromatography \\ LC_{50} & lethal concentration, 50\% \ kill \\ LC_{Lo} & lethal concentration, low \\ LD_{50} & lethal dose, 50\% \ kill \\ LD_{Lo} & lethal dose, low \\ LDH & lactic dehydrogenase \\ LH & luteinizing hormone \\ \end{array}$

LOAEL lowest-observed-adverse-effect level LSE Levels of Significant Exposure

LT₅₀ lethal time, 50% kill

m meter

MA trans,trans-muconic acid MAL maximum allowable level

mCi millicurie

MCL maximum contaminant level

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MCLG maximum contaminant level goal

MF modifying factor MFO mixed function oxidase

mg milligram mL milliliter mm millimeter

mmHg millimeters of mercury

mmol millimole

mppcf millions of particles per cubic foot

MRL Minimal Risk Level MS mass spectrometry

mt metric ton

NAAQS National Ambient Air Quality Standard

NAS National Academy of Science

NATICH National Air Toxics Information Clearinghouse

NATO North Atlantic Treaty Organization NCE normochromatic erythrocytes

NCEH National Center for Environmental Health

NCI National Cancer Institute

ND not detected

NFPA National Fire Protection Association

ng nanogram

NHANES National Health and Nutrition Examination Survey
NIEHS National Institute of Environmental Health Sciences
NIOSH National Institute for Occupational Safety and Health
NIOSHTIC NIOSH's Computerized Information Retrieval System

NLM National Library of Medicine

nm nanometer nmol nanomole

NOAEL no-observed-adverse-effect level

NOES National Occupational Exposure Survey NOHS National Occupational Hazard Survey

NPD nitrogen phosphorus detection

NPDES National Pollutant Discharge Elimination System

NPL National Priorities List

NR not reported

NRC National Research Council

NS not specified

NSPS New Source Performance Standards NTIS National Technical Information Service

NTP National Toxicology Program ODW Office of Drinking Water, EPA

OERR Office of Emergency and Remedial Response, EPA

OHM/TADS Oil and Hazardous Materials/Technical Assistance Data System

OPP Office of Pesticide Programs, EPA

OPPT Office of Pollution Prevention and Toxics, EPA

OPPTS Office of Prevention, Pesticides and Toxic Substances, EPA

OR odds ratio

OSHA Occupational Safety and Health Administration

OSW Office of Solid Waste, EPA OTS Office of Toxic Substances

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OW Office of Water

OWRS Office of Water Regulations and Standards, EPA

PAH polycyclic aromatic hydrocarbon

PBPD physiologically based pharmacodynamic PBPK physiologically based pharmacokinetic

PCE polychromatic erythrocytes PEL permissible exposure limit

PEL-C permissible exposure limit-ceiling value

pg picogram

PHS Public Health Service
PID photo ionization detector

pmol picomole

PMR proportionate mortality ratio

ppb parts per billion ppm parts per million ppt parts per trillion

PSNS pretreatment standards for new sources

RBC red blood cell

REL recommended exposure level/limit

REL-C recommended exposure level-ceiling value RfC reference concentration (inhalation)

RfD reference dose (oral)
RNA ribonucleic acid
RQ reportable quantity

RTECS Registry of Toxic Effects of Chemical Substances SARA Superfund Amendments and Reauthorization Act

SCE sister chromatid exchange

SGOT serum glutamic oxaloacetic transaminase (same as aspartate aminotransferase or AST)
SGPT serum glutamic pyruvic transaminase (same as alanine aminotransferase or ALT)

SIC standard industrial classification

SIM selected ion monitoring

SMCL secondary maximum contaminant level

SMR standardized mortality ratio

SNARL suggested no adverse response level

SPEGL Short-Term Public Emergency Guidance Level

STEL short term exposure limit STORET Storage and Retrieval

TD₅₀ toxic dose, 50% specific toxic effect

TLV threshold limit value

TLV-C threshold limit value-ceiling value

TOC total organic carbon

TPQ threshold planning quantity
TRI Toxics Release Inventory
TSCA Toxic Substances Control Act

TWA time-weighted average UF uncertainty factor U.S. United States

USDA United States Department of Agriculture

USGS United States Geological Survey VOC volatile organic compound

WBC white blood cell

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WHO World Health Organization

| > | greater than |
|---------------|--------------------------|
| \geq | greater than or equal to |
| = | equal to |
| < | less than |
| > = < < < < % | less than or equal to |
| % | percent |
| α | alpha |
| β | beta |
| γ | gamma |
| δ | delta |
| μm | micrometer |
| μg | microgram |
| q_1^* | cancer slope factor |
| _ | negative |
| + | positive |
| (+) | weakly positive result |
| (-) | weakly negative result |