

6. POTENTIAL FOR HUMAN EXPOSURE

6.1 OVERVIEW

Barium has been identified in at least 798 of the 1,684 hazardous waste sites that have been proposed for inclusion on the EPA National Priorities List (NPL) (HazDat 2006). However, the number of sites evaluated for barium is not known. The frequency of these sites can be seen in Figure 6-1. Of these sites, 794 are located within the United States, 1 is in the Territory of Guam, and 3 are located in the Commonwealth of Puerto Rico (the Territory of Guam and the Commonwealth of Puerto Rico are not shown).

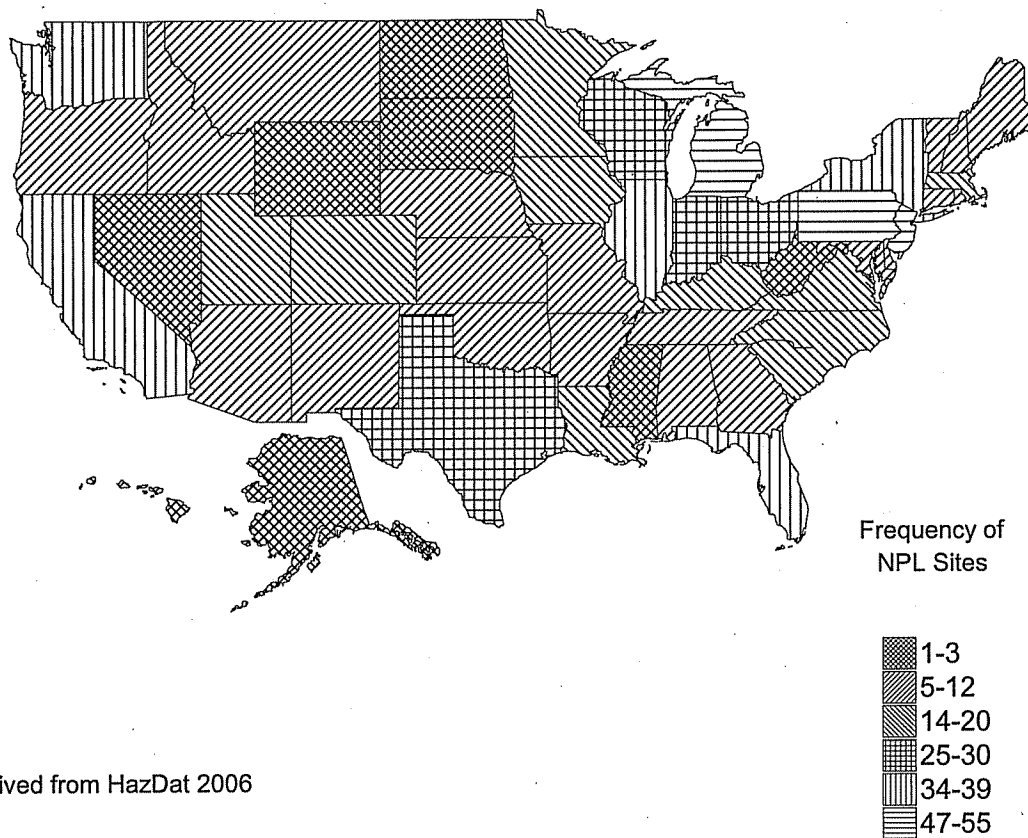
Barium is a naturally occurring component of minerals that are found in small but widely distributed amounts in the earth's crust, especially in igneous rocks, sandstone, shale, and coal (Kunesh 1978; Miner 1969a). Barium enters the environment naturally through the weathering of rocks and minerals. Anthropogenic releases are primarily associated with industrial processes. Barium is present in the atmosphere, urban and rural surface water, soils, and many foods.

Under natural conditions, barium is stable in the +2 valence state and is found primarily in the form of inorganic complexes. Conditions such as pH, Eh (oxidation-reduction potential), cation exchange capacity, and the presence of sulfate, carbonate, and metal oxides (e.g., oxides of aluminum, manganese, silicon, and titanium) will affect the partitioning of barium and its compounds in the environment. The major features of the biogeochemical cycle of barium include wet and dry deposition to land and surface water, leaching from geological formations to groundwater, adsorption to soil and sediment particulates, and biomagnification in terrestrial and aquatic food chains.

The general population is exposed to barium through consumption of drinking water and foods, usually at low levels. Workers in barium mining or processing industries and individuals who reside near such industries might be exposed to relatively high levels, primarily through the inhalation of fugitive dust containing barium compounds. The most recent occupational exposure estimates indicate that about 10,000 people were potentially exposed to barium and about 474,000 to barium compounds in workplace environments in the United States in 1980 (NIOSH 1989a).

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Figure 6-1. Frequency of NPL Sites with Barium Contamination



Derived from HazDat 2006

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6.2 RELEASES TO THE ENVIRONMENT

The Toxics Release Inventory (TRI) data should be used with caution because only certain types of facilities are required to report (EPA 2005d). This is not an exhaustive list. Manufacturing and processing facilities are required to report information to the TRI only if they employ 10 or more full-time employees; if their facility is included in Standard Industrial Classification (SIC) Codes 10 (except 1011, 1081, and 1094), 12 (except 1241), 20–39, 4911 (limited to facilities that combust coal and/or oil for the purpose of generating electricity for distribution in commerce), 4931 (limited to facilities that combust coal and/or oil for the purpose of generating electricity for distribution in commerce), 4939 (limited to facilities that combust coal and/or oil for the purpose of generating electricity for distribution in commerce), 4953 (limited to facilities regulated under RCRA Subtitle C, 42 U.S.C. section 6921 et seq.), 5169, 5171, and 7389 (limited S.C. section 6921 et seq.), 5169, 5171, and 7389 (limited to facilities primarily engaged in solvents recovery services on a contract or fee basis); and if their facility produces, imports, or processes $\geq 25,000$ pounds of any TRI chemical or otherwise uses $>10,000$ pounds of a TRI chemical in a calendar year (EPA 2005d).

Barium is a highly reactive metal that occurs naturally only in a combined state. The element is released to environmental media by both natural processes and anthropogenic sources.

According to the SARA Section 313 Toxics Release Inventory (TRI), an estimated total of 230 million pounds (105,000 metric tons) of barium and barium compounds were released to the environment from manufacturing and processing facilities in the United States in 2004 (TRI04 2006) (see Tables 6-1 and 6-2). Most of these barium releases were to land. The TRI data must be viewed with caution since only certain types of facilities were required to report. This is not an exhaustive list.

6.2.1 Air

Estimated combined releases of 2.51 million pounds (1,140 metric tons) of barium (0.35 million pounds) and barium compounds (2.16 million pounds) to the atmosphere from 1,107 domestic manufacturing and processing facilities in 2004, accounted for about 1.09% of the estimated total environmental releases from facilities required to report to the TRI (TRI04 2006). These releases are summarized in Tables 6-1 and 6-2.

Barium is released primarily to the atmosphere as a result of industrial emissions during the mining, refining, and production of barium and barium chemicals, fossil fuel combustion (Miner 1969a), and

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Table 6-1. Releases to the Environment from Facilities that Produce, Process, or Use Barium^a

State ^c	RF ^d	Reported amounts released in pounds per year ^b							Total release	
		Air ^e	Water ^f	UI ^g	Land ^h	Other ⁱ	On-site ^j	Off-site ^k	On- and off-site	
AK	3	24,620	0	0	306,812	0	146,620	184,812	331,432	
AL	5	6,963	504	0	39,889	4,748	7,459	44,645	52,104	
AZ	2	1,062	0	0	1,950,946	0	1,823,807	128,201	1,952,008	
CA	2	1,911	0	0	797,507	78	799,418	78	799,496	
CO	1	0	0	No data	0	0	No data	0	0	
CT	1	4	4	0	0	3,515	4	3,519	3,523	
DE	1	20	64	0	0	0	84	0	84	
GA	4	81	0	0	3,757	0	81	3,757	3,837	
IA	4	36,228	0	0	163	0	36,228	163	36,391	
ID	1	14	0	0	130,611	0	130,625	0	130,625	
IL	3	9,428	61	0	45,553	0	9,489	45,553	55,042	
IN	2	10	255	0	18,074	16,900	10	35,229	35,239	
KS	2	7,501	0	0	161,964	526,878	169,465	526,878	696,343	
KY	2	75,258	0	0	0	0	75,258	0	75,258	
MI	5	230	666	0	44,245	271,175	896	315,420	316,316	
MN	2	114,719	0	0	694	0	114,719	694	115,413	
MO	1	0	0	0	0	81	0	81	81	
NC	3	11	0	0	1,559	0	11	1,559	1,570	
NE	6	23,979	3,320	0	37,786	362,667	65,080	362,672	427,752	
NJ	2	89	0	0	30	272	89	302	391	
NM	1	0	0	0	0	0	0	0	0	
NV	2	1,243	0	0	817,749	0	818,992	0	818,992	
NY	5	35,001	137	0	3,073	15,991	35,140	19,062	54,202	
OH	14	866	372	16,649	287,615	6,472	220,822	91,152	311,974	
OK	1	1,906	5	0	250	0	1,906	255	2,161	
OR	1	0	0	0	230,293	1	230,293	1	230,294	
PA	4	756	0	0	344,131	51,746	322,501	74,132	396,633	
SC	2	253	0	0	0	72	253	72	325	
SD	1	500	0	0	39,480	0	39,980	0	39,980	
TN	1	0	0	No data	0	0	No data	0	0	
TX	9	7,392	4,029	0	71,083	7	81,473	1,038	82,511	
UT	2	10	0	0	31,035	0	31,010	35	31,045	
VA	2	0	1,900	0	111,900	0	95,900	17,900	113,800	
WI	1	0	0	0	257,400	0	0	257,400	257,400	

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Table 6-1. Releases to the Environment from Facilities that Produce, Process, or Use Barium^a

State ^c	RF ^d	Reported amounts released in pounds per year ^b					Total release		
		Air ^e	Water ^f	UI ^g	Land ^h	Other ⁱ	On-site ^j	Off-site ^k	On- and off-site
WV	2	4,214	10	0	77,821	201	82,045	202	82,246
Total	100	354,269	11,327	16,649	5,811,421	1,260,803	5,339,658	2,114,811	7,454,470

^aThe TRI data should be used with caution since only certain types of facilities are required to report. This is not an exhaustive list. Data are rounded to nearest whole number.

^bData in TRI are maximum amounts released by each facility.

^cPost office state abbreviations are used.

^dNumber of reporting facilities.

^eThe sum of fugitive and point source releases are included in releases to air by a given facility.

^fSurface water discharges, waste water treatment-(metals only), and publicly owned treatment works (POTWs) (metal and metal compounds).

^gClass I wells, Class II-V wells, and underground injection.

^hResource Conservation and Recovery Act (RCRA) subtitle C landfills; other on-site landfills, land treatment, surface impoundments, other land disposal, other landfills.

ⁱStorage only, solidification/stabilization (metals only), other off-site management, transfers to waste broker for disposal, unknown

^jThe sum of all releases of the chemical to air, land, water, and underground injection wells.

^kTotal amount of chemical transferred off-site, including to POTWs.

RF = reporting facilities; UI = underground injection

Source: TRI04 2006 (Data are from 2004)

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Table 6-2. Releases to the Environment from Facilities that Produce, Process, or Use Barium Compounds^a

State ^c	RF ^d	Reported amounts released in pounds per year ^b							Total release	
		Air ^e	Water ^f	UI ^g	Land ^h	Other ⁱ	On-site ^j	Off-site ^k	On- and off-site	
AK	1	16,122	0	0	16,122	0	32,244	0	32,244	
AL	34	69,621	94,049	3,900	11,966,847	67,278	12,062,542	139,153	12,201,695	
AR	16	28,134	107,680	0	2,906,741	44,565	3,038,693	48,427	3,087,120	
AZ	10	11,779	0	0	3,951,882	61	3,962,437	1,285	3,963,722	
CA	14	8,998	514	0	56,003	3,624	20,826	48,313	69,139	
CO	18	14,440	2,232	0	8,687,234	7,500	5,019,424	3,691,982	8,711,406	
CT	4	505	5	0	66,963	22,315	540	89,248	89,788	
DE	3	17,256	9,034	0	480,594	121,025	486,290	141,619	627,909	
FL	22	49,588	9,750	0	2,232,733	216,852	2,134,494	374,429	2,508,924	
GA	29	61,732	112,867	0	8,660,091	19,737	8,798,097	56,330	8,854,427	
HI	1	45	0	0	29,331	0	45	29,331	29,376	
IA	24	142,028	3,709	0	4,358,835	229,865	3,293,915	1,440,522	4,734,437	
ID	6	8,334	11,400	0	137,167	83	98,901	58,083	156,984	
IL	56	218,005	114,459	0	13,593,791	370,276	6,061,909	8,234,621	14,296,530	
IN	45	72,823	31,471	0	8,274,849	82,196	7,095,068	1,366,271	8,461,340	
KS	12	78,277	938	0	4,774,271	250	4,853,486	250	4,853,736	
KY	38	49,663	76,268	0	7,134,938	211,117	5,405,100	2,066,886	7,471,986	
LA	17	70,200	40,931	1,367	4,315,727	8,307	4,411,344	25,188	4,436,532	
MA	14	2,275	1,283	0	257,892	21,982	20,043	263,389	283,432	
MD	18	9,340	1,880	59	468,562	697,038	90,155	1,086,724	1,176,879	
ME	6	1,301	4,100	0	81,692	0	71,193	15,900	87,093	
MI	32	55,968	125,134	56	10,513,385	24,961	7,965,087	2,754,417	10,719,504	
MN	22	48,772	22,540	0	8,233,091	71,522	7,430,251	945,673	8,375,924	
MO	30	231,257	11,722	0	10,226,716	731	10,356,850	113,576	10,470,426	
MS	16	5,973	14,703	0	1,626,877	528	1,627,878	20,203	1,648,081	
MT	6	111,530	781	0	8,892,216	175,814	8,951,015	229,325	9,180,340	
NC	31	30,632	68,807	0	3,082,682	171,007	2,921,804	431,324	3,353,127	
ND	9	39,926	24,052	0	13,826,846	6,786	7,126,843	6,770,767	13,897,610	
NE	11	37,560	52	0	3,982,921	6,350	3,739,570	287,313	4,026,883	
NH	5	1,532	0	0	26,500	1,583	9,632	19,983	29,615	
NJ	20	7,826	9,949	0	96,325	297,053	7,832	403,321	411,153	
NM	6	13,400	250	0	5,210,450	750	5,204,409	20,441	5,224,850	
NV	4	25,974	0	0	1,432,649	37	1,458,623	37	1,458,660	
NY	24	18,085	136,769	0	1,454,199	311,766	790,432	1,130,387	1,920,819	
OH	94	43,103	82,835	319	8,642,836	424,109	5,383,148	3,810,054	9,193,202	
OK	12	25,481	11,600	0	2,005,538	19	1,737,944	304,694	2,042,638	
OR	5	12,393	2,120	0	115,146	0	100,513	29,146	129,659	

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Table 6-2. Releases to the Environment from Facilities that Produce, Process, or Use Barium Compounds^a

State ^c	RF ^d	Air ^e	Reported amounts released in pounds per year ^b							Total release	
			Water ^f	UI ^g	Land ^h	Other ⁱ	On-site ^j	Off-site ^k	On- and off-site		
PA	72	50,576	40,857	250	3,948,212	422,230	1,469,982	2,992,143	4,462,125		
PR	1	1,542	0	0	91	0	1,633	0	1,633		
RI	3	5	23	0	1,072	477	28	1,549	1,577		
SC	36	27,220	51,763	0	1,313,342	678,479	1,263,992	806,812	2,070,803		
SD	2	1,057	36	0	731,856	0	681,949	51,000	732,949		
TN	24	64,154	106,884	0	4,653,790	334	4,240,684	584,478	4,825,162		
TX	48	158,316	67,044	0	16,122,300	3,011,576	16,164,215	3,195,022	19,359,236		
UT	9	5,422	100	0	3,661,314	3,902	3,510,292	160,446	3,670,738		
VA	28	25,919	26,448	0	2,006,961	189,025	1,704,517	543,836	2,248,353		
VT	1	250	5	0	0	28,667	250	28,672	28,922		
WA	11	1,318	3,118	0	2,274,705	42	2,119,156	160,027	2,279,183		
WI	30	37,077	20,145	0	890,567	1,023,144	342,849	1,628,084	1,970,933		
WV	20	75,248	18,436	0	5,243,526	78,000	4,479,752	935,458	5,415,210		
WY	7	68,528	3,229	0	7,343,092	0	6,751,399	663,450	7,414,849		
Total	1,007	2,156,511	1,471,972	5,951	210,011,467	9,052,963	174,499,278	48,199,586	222,698,864		

^aThe TRI data should be used with caution since only certain types of facilities are required to report. This is not an exhaustive list. Data are rounded to nearest whole number.

^bData in TRI are maximum amounts released by each facility.

^cPost office state abbreviations are used.

^dNumber of reporting facilities.

^eThe sum of fugitive and point source releases are included in releases to air by a given facility.

^fSurface water discharges, waste water treatment-(metals only), and publicly owned treatment works (POTWs) (metal and metal compounds).

^gClass I wells, Class II-V wells, and underground injection.

^hResource Conservation and Recovery Act (RCRA) subtitle C landfills; other on-site landfills, land treatment, surface impoundments, other land disposal, other landfills.

ⁱStorage only, solidification/stabilization (metals only), other off-site management, transfers to waste broker for disposal, unknown

^jThe sum of all releases of the chemical to air, land, water, and underground injection wells.

^kTotal amount of chemical transferred off-site, including to POTWs.

RF = reporting facilities; UI = underground injection

Source: TRI04 2006 (Data are from 2004)

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entrainment of soil and rock dust into the air (Schroeder 1970). In addition, coal ash, containing widely variable amounts of barium, is also a source of airborne barium particulates (Miner 1969a; Schroeder 1970). In 1969, an estimated 18% of the total U.S. barium emissions to the atmosphere resulted from the processing of barite ore, and more than 28% of the total was estimated to be from the production of barium chemicals. The manufacture of various end products (e.g., drilling well muds, and glass, paint, and rubber products) and the combustion of coal were estimated to account for an additional 23 and 26% of the total barium emissions for 1969, respectively (Davis 1972).

Estimates of barium releases from individual industrial processes are available for particulate emissions from the drying and calcining of barium compounds and for fugitive dust emissions during the processing of barite ore. Soluble barium compounds (unspecified) are emitted as particulates from barium chemical dryers and calciners to the atmosphere during the processing of barium carbonate, barium chloride, and barium hydroxide (Reznik and Toy 1978). Uncontrolled particulate emissions of soluble barium compounds from chemical dryers and calciners during barium processing operations may range from 0.04 to 10 g/kg of final product. Controlled particulate emissions are less than 0.25 g/kg of final product. Based on an uncontrolled emission factor of 5 g/kg and a controlled emission factor of 0.25 g/kg, total particulate emissions from the drying and calcining of barium carbonate, barium chloride, and barium hydroxide are estimated to be 160 metric tons (352,800 pounds) per year (Reznik and Toy 1978).

Fugitive dust emissions occur during processing (grinding and mixing) of barite ore and may also occur during the loading of bulk product of various barium compounds into railroad hopper cars (Reznik and Toy 1978). Based on an emission factor of 1 g/kg, total emissions of fugitive dust from the domestic barium chemicals industry during the grinding of barite ore have been estimated to be approximately 90 metric tons (198,450 pounds) per year (Reznik and Toy 1978). Other particulate emissions from the industrial production of barium compounds include an estimated 820 metric tons (1.8 million pounds) per year from uncontrolled kilns during the processing of barite ore and 8 metric tons (17,640 pounds) per year from black ash (i.e., barium sulfide) rotary kilns during the production of barium hydroxide (Reznik and Toy 1978). Electric utilities that burn bituminous coal emit a small fraction of the barium contained in coal into the air. For example, it is estimated that 830 pounds/year of barium are released to air from a 650 megawatt (MW) plant, in comparison to 270,000 pounds/year released as ash to land-based waste sites (Rubin 1999).

The use of barium in the form of organometallic compounds as a smoke suppressant in diesel fuels results in the release of solids to the atmosphere (Miner 1969a; Ng and Patterson 1982; Schroeder 1970). The

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maximum concentration of soluble barium in exhaust gases containing barium-based smoke suppressants released from test diesel engines and operating diesel vehicles is estimated to be $12,000 \mu\text{g}/\text{m}^3$, when the barium concentration in the diesel fuel is 0.075% by weight and 25% of the exhausted barium (at a sampling point 10 feet from the engine and upstream from the muffler) is soluble (Golothan 1967). Thus, 1 L of this exhaust gas contains an estimated 12 μg soluble barium or 48 μg total barium (Schroeder 1970). However, recent legislation requiring the use of low-sulfur fuel in diesel engines has eliminated the need for barium as a sulfur-scavenging additive and, therefore, has greatly reduced the emissions of barium from diesel engine exhaust (Schauer et al. 1999; Winkler 2002).

6.2.2 Water

Estimated combined releases of 1.48 million pounds (674 metric tons) of barium (0.01 million pounds) and barium compounds (1.47 million pounds) to surface water from 1,107 domestic manufacturing and processing facilities in 2004, accounted for about 0.64% of the estimated total environmental releases from facilities required to report to the TRI (TRI04 2006).

The primary source of naturally occurring barium in drinking water results from the leaching and eroding of sedimentary rocks into groundwater (Kojola et al. 1978). Although barium occurs naturally in most surface water bodies (i.e., approximately 99% of those examined) (DOI 1970), releases of barium to surface waters from natural sources are much lower than those to groundwater (Kojola et al. 1978).

About 80% of the barium produced is used as barite to make high-density oil and gas well drilling muds, and during offshore drilling operations there are periodic discharges of drilling wastes in the form of cuttings and muds into the ocean (Ng and Patterson 1982). For example, in the Santa Barbara Channel region, about 10% of the muds used are lost into the ocean (Ng and Patterson 1982). Operations involving three drilling platforms in the Santa Maria Basin off the coast of central California released approximately 1.8×10^6 kg of barium to the ocean in discharged muds, cuttings, and waste water from 1986 to 1994 (Phillips et al. 1998). The use of barium in offshore drilling operations may increase barium pollution, especially in coastal sediments (Ng and Patterson 1982).

6.2.3 Soil

Estimated combined releases of 216 million pounds (98,095 metric tons) of barium (5.81 million pounds) and barium compounds (210 million pounds) to soils from 1,107 domestic manufacturing and processing facilities in 2004, accounted for about 93.7% of the estimated total environmental releases from facilities

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required to report to the TRI (TRI04 2006). An additional combined total of 0.023 million pounds (10 metric tons) from barium (0.017 million pounds) and barium compounds (0.006 million pounds), constituting about 0.01% of the total environmental emissions, were released via underground injection (TRI04 2006). These releases are summarized in Tables 6-1 and 6-2.

The process of drilling for crude oil and natural gas generates waste drilling fluids or muds, which are often disposed of by land farming. Most of these fluids are water based and contain barite and other metal salts. Thus, barium may be introduced into soils as the result of land farming these slurried reserve pit wastes (Bates 1988).

The use of barium fluorosilicate and carbonate as insecticides (Beliles 1979; Meister 2004) might also contribute to the presence of barium in agricultural soils.

Barium has been detected with a positive geometric mean concentration of 100.5 ppm in soil samples from approximately 52% of the hazardous waste sites that have had samples analyzed by the CLP (CLPSD 1989). Note that these data from the CLPSD represent frequency of occurrence and concentration data for NPL sites only.

6.3 ENVIRONMENTAL FATE

6.3.1 Transport and Partitioning

Most barium released to the environment from industrial sources is in forms that do not become widely dispersed (Ng and Patterson 1982). In the atmosphere, barium is likely to be present in particulate form (EPA 1984). Although chemical reactions may cause changes in speciation of barium in air, the main mechanisms for the removal of barium compounds from the atmosphere are likely to be wet and dry deposition (EPA 1984).

In aquatic media, barium is likely to precipitate out of solution as an insoluble salt (i.e., as BaSO_4 or BaCO_3). Waterborne barium may also adsorb to suspended particulate matter through the formation of ion pairs with natural anions such as bicarbonate or sulfate in the matter (Bodek et al. 1988; EPA 1984; Giusti et al. 1993; Lagas et al. 1984; Tanizaki et al. 1992). Precipitation of barium sulfate salts is accelerated when rivers enter the ocean because of the high sulfate content (905 mg/L) in the ocean (Bowen 1966; WHO 2001). It is estimated that only 0.006% of the total barium input into oceans from freshwater sources remains in solution (Chow et al. 1978; WHO 2001). Sedimentation of suspended

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solids removes a large portion of the barium content from surface waters (Benes et al. 1983). There is evidence to suggest that the precipitation of barium from the surface of fresh and marine waters occurs, in part, as the result of the barite crystal formation in microorganisms (González-Muñoz et al. 2003).

Barium in sediments is found largely in the form of barium sulfate (barite). Coarse silt sediment in a turbulent environment will often grind and cleave the barium sulfate from the sediment particles leaving a buildup of dense barites (Merefield 1987). Estimated soil:water distribution coefficients (K_d) (i.e., the ratio of the quantity of barium sorbed per gram of sorbent to the concentration of barium remaining in solution at equilibrium) range from 200 to 2,800 for sediments and sandy loam soils (DOE 1984; Rai et al. 1984).

The uptake of barium by fish and marine organisms is also an important removal mechanism (Bowen 1966; Schroeder 1970). Barium levels in sea water range from 2 to 63 $\mu\text{g/L}$ with a mean concentration of about 13 $\mu\text{g/L}$ (Bowen 1979). Barium was found to bioconcentrate in marine plants by a factor of 400–4,000 times the level present in the water (Bowen 1966). Bioconcentration factors in marine animals, plankton, and brown algae of 100, 120, and 260, respectively, have been reported (Schroeder 1970). In freshwater, a bioconcentration factor of 129 was estimated in fish where the barium in water was 0.07 mg/L (Hope et al. 1996).

Barium added to soils (e.g., from the land farming of waste drilling muds) may either be taken up by vegetation or transported through soil with precipitation (Bates 1988). Relative to the amount of barium found in soils, little is typically bioconcentrated by plants (Schroeder 1970). For example, a bioconcentration factor of 0.4 has been estimated for plants in a Virginia floodplain with a barium soil concentration of 104.2 mg/kg (Hope et al. 1996). However, there are some plants, such as legumes, forage plants, Brazil nuts, and mushrooms that accumulate barium (Aruguete et al. 1998; IPCS 1991; WHO 2001). Bioconcentration factors from 2 to 20 have been reported for tomatoes and soybeans (WHO 2001).

Barium is not very mobile in most soil systems, due to the formation of water-insoluble salts and an inability of the barium ion to form soluble complexes with fulvic and humic acids (WHO 2001). The rate of transportation of barium in soil is dependent on the characteristics of the soil material. Soil properties that influence the transportation of barium to groundwater are cation exchange capacity, calcium carbonate (CaCO_3) content and pH. In soil with a high cation exchange capacity (e.g., fine textured mineral soils or soils with high organic matter content), barium mobility will be limited by adsorption (Bates 1988; Kabata-Pendias and Pendias 1984). High CaCO_3 content limits mobility by precipitation of

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the element as BaCO_3 (Lagas et al. 1984). Barium will also precipitate as barium sulfate in the presence of sulfate ions (Bodek et al. 1988; Lagas et al. 1984). Barium is more mobile and is more likely to be leached from soils in the presence of chloride due to the high solubility of barium chloride as compared to other chemical forms of barium (Bates 1988; Lagas et al. 1984). Barium may become more mobile in soils under acid conditions as barium in water-insoluble salts, such as barium sulfate and carbonate, becomes more soluble (WHO 2001). Barium complexes with fatty acids (e.g., in acidic landfill leachate) will be much more mobile in the soil due to the lower charge of these complexes and subsequent reduction in adsorption capacity (Lagas et al. 1984).

Barium mobility in soil is reduced by the precipitation of barium carbonate and sulfate. Humic and fulvic acid have not been found to increase the mobility of barium (EPA 1984).

6.3.2 Transformation and Degradation

6.3.2.1 Air

Elemental barium undergoes oxidation in air and is oxidized readily in moist air (Boffito 2002; EPA 1983; Kresse et al. 2007; Kunesh 1978). The residence time of barium in the atmosphere may be several days, depending on the size of the particulate formed, the chemical nature of the particulate, and environmental factors such as rainfall (EPA 1984; WHO 2001).

6.3.2.2 Water

Under natural conditions, barium will form compounds in the +2 oxidation state. Barium does not hydrolyze appreciably except in highly alkaline environments (i.e., at pH levels ≥ 10) (Bodek et al. 1988).

Appreciable levels of barium sulfate occur because natural water often contains high sulfate concentrations, especially ocean water. Since the solubility of barium sulfate is low, only trace amounts of barium dissolve in surface water (Bodek et al. 1988; NAS 1977). At pH levels of 9.3 or below, barium sulfate may limit barium concentrations in natural waters (Bodek et al. 1988). The solubility of barium sulfate increases considerably in the presence of chloride (Cl^-) and other anions (e.g., NO_3^- and CO_3^{2-}), and at pH levels of 9.3 or below, the barium ion (Ba^{2+}) is the dominant species (Bodek et al. 1988; NAS 1977). The Ba^{2+} ion is stable under the pH-Eh range of natural systems. However, natural and treated waters usually contain sufficient sulfate so that a barium ion concentration of more than 1,000–

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1,500 µg/L cannot be maintained in solution (EPA 1983; Hem 1959; Lagas et al. 1984; McCabe et al. 1970).

As pH levels increase above 9.3 and in the presence of carbonate, barium carbonate becomes the dominant species (Bodek et al. 1988; Singer 1974). Barium carbonate also exhibits fast precipitation kinetics and very low solubility and in alkaline environments limits the soluble barium concentration (Faust and Aly 1981; Hem 1959; Rai et al. 1984; Singer 1974). Barium forms salts of low solubility with arsenate, chromate, fluoride, oxalate, and phosphate ions (Bodek et al. 1988; EPA 1983; Kunesh 1978). The chloride, hydroxide, and nitrate of barium are water-soluble (Bodek et al. 1988; EPA 1983; Kirkpatrick 1978) and are frequently detected in aqueous environments (Rai et al. 1984).

Barium also forms complexes with natural organics in water (e.g., fatty acids in acidic landfill leachates) to a limited extent (Lagas et al. 1984; Morel 1983; Rai et al. 1984).

6.3.2.3 Sediment and Soil

Barium reacts with metal oxides and hydroxides in soil and is subsequently adsorbed onto soil particulates (Hem 1959; Rai et al. 1984). Adsorption onto metal oxides in soils and sediments probably acts as a control over the concentration of barium in natural waters (Bodek et al. 1988). Under typical environmental conditions, barium displaces other adsorbed alkaline earth metals from MnO_2 , SiO_2 , and TiO_2 (Rai et al. 1984). However, barium is displaced from Al_2O_3 by other alkaline earth metals (Rai et al. 1984). The ionic radius of the barium $2+$ ion, its typical oxidation state, makes isomorphous substitution possible only with strontium, and generally not with the other members of the alkaline earth elements (Kirkpatrick 1978). Among the other elements that occur with barium in nature, substitution is common only with potassium but not with the smaller ions of sodium, iron, manganese, aluminum, and silicon (Kirkpatrick 1978).

Barium is also adsorbed onto soil and subsoil through electrostatic interactions (Bodek et al. 1988; Singer 1974). The cation exchange capacity of the sorbent largely controls the retention of barium in soils (Bodek et al. 1988). Barium is strongly adsorbed by clay minerals (Kabata-Pendias and Pendias 1984; Lagas et al. 1984).

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Barium can also form salts with acetate, nitrate, chloride, and hydroxide ions in soil. The mobility of barium in soils increases upon formation of these water soluble salts (Bodek et al. 1988). In general, the solubility of barium compounds increases with decreasing pH.

6.4 LEVELS MONITORED OR ESTIMATED IN THE ENVIRONMENT

Reliable evaluation of the potential for human exposure to barium depends in part on the reliability of supporting analytical data from environmental samples and biological specimens. Concentrations of barium in unpolluted atmospheres and in pristine surface waters are often so low as to be near the limits of current analytical methods. In reviewing data on barium levels monitored or estimated in the environment, it should also be noted that the amount of chemical identified analytically is not necessarily equivalent to the amount that is bioavailable. The analytical methods available for monitoring barium in a variety of environmental media are detailed in Chapter 7.

6.4.1 Air

The concentration of barium in ambient air is estimated to be $<0.05 \mu\text{g}/\text{m}^3$ (IPCS 1991). Airborne barium likely exists as the carbonate or sulfate salts and is carried on particulate matter that results as a consequence of natural processes (e.g., suspension of soil dust) or anthropogenic activities (e.g., combustion process, mining and calcining of barium ores) (WHO 2001). However, there is no apparent correlation between the degree of industrialization and barium concentrations in ambient air (Winkler 2002). Particulate matter from diesel exhaust was once a source of barium in ambient air. However, barium emissions from diesel engines has been greatly diminished to near zero emissions with the current use of low-sulfur diesel fuels, which do not require the addition of barium as a sulfur-scavenging agent (Hildemann et al. 1991; Schauer et al. 1999; Shahin et al. 2000; Winkler 2002).

Tabor and Warren (1958) report urban and suburban air concentrations of barium ranging from <0.005 to $1.5 \mu\text{g}/\text{m}^3$. In another study of barium concentrations in ambient air, values ranged from 0.0015 to $0.95 \text{ mg}/\text{m}^3$ (EPA 1984). No distinct pattern related to industrialization appeared in the results reported on 754 samples from 18 cities and four suburban areas in the United States. For example, in Houston, Texas and its suburbs, 76% of the samples contained barium at levels ranging from 0.005 to $1.5 \mu\text{g}/\text{m}^3$, whereas in Fort Worth, Texas, 66% of the samples had values $<0.005 \mu\text{g}/\text{m}^3$ (Tabor and Warren 1958).

Another compilation of atmospheric data shows barium concentrations in urban atmospheres of North America ranging from 2×10^{-4} to $2.8 \times 10^{-2} \mu\text{g}/\text{m}^3$ with a mean concentration of $1.2 \times 10^{-2} \mu\text{g}/\text{m}^3$ (Bowen

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1979). In contrast, barium levels in samples from the South Pole and northern Norway were 1.6×10^{-5} and $7.3 \times 10^{-4} \mu\text{g}/\text{m}^3$, respectively (Bowen 1979). Mean barium concentrations in background air collected between April and October 2002 on the campus of the University of Birmingham, United Kingdom, were 0.32 and $1.4 \text{ ng}/\text{m}^3$ in the <0.5 and $3.0\text{--}7.2 \mu\text{m}$ particulate matter fractions, respectively (Birmili et al. 2006).

Maximum ground-level barium concentrations (as soluble compounds) associated with uncontrolled atmospheric particulate emissions from chemical dryers and calciners at barium-processing plants have been estimated (using dispersion modeling) to range from 1.3 to $330 \mu\text{g}/\text{m}^3$ over a 24-hour averaging time at locations along facility boundaries (i.e., away from the source of emission) (Reznik and Toy 1978).

Barium has been measured in dust samples taken from 49 residences in Ottawa, Canada. Mean and median concentrations of 405.56 and 222.22 mg barium/kg dust, respectively, were measured within a sub-fraction of the dust samples where the particulate sizes ranged from 100 to $250 \mu\text{m}$ (Butte and Heinzow 2002; Rasmussen et al. 2001).

Barium has been measured in rain and snow collected near Claremont, New Hampshire in 1996–1997 (Feng et al. 2000). Barium concentrations in rain ranged from 0.22 to $0.84 \mu\text{g}/\text{L}$ with a mean concentration of $0.39 \mu\text{g}/\text{L}$. In snow, barium concentrations ranged from 0.64 to $7.44 \mu\text{g}/\text{L}$ with a mean concentration of $1.5 \mu\text{g}/\text{L}$.

Barium has been detected in air samples collected at 24 of the 798 hazardous waste sites where barium has been detected in some environmental medium (HazDat 2006). The HazDat information includes data from both NPL and other Superfund sites. Concentrations of barium in air ranged from 0.015 to $327,000,000 \mu\text{g}/\text{m}^3$ in 16 onsite (HazDat 2006). In comparison, concentrations of barium in air ranged from 0.0135 to $561,000,000 \mu\text{g}/\text{m}^3$ in 12 offsite samples (HazDat 2006).

6.4.2 Water

Barium has been found in almost all raw surface waters and public drinking water supplies sampled (i.e., approximately 99%) (Kopp 1969) at concentrations ranging from ≤ 5 to $15,000 \mu\text{g}/\text{L}$ with mean concentrations generally on the order of $10\text{--}60 \mu\text{g}/\text{L}$ (Barnett et al. 1969; Bowen 1979; DOI 1970; Durfor and Becker 1964; Durum and Haffty 1961; Elinder and Zenz 1994; EPA 2005c; Kopp 1969; Longerich et al. 1991; McCabe et al. 1970; Neal et al. 1996; Saleh and Wilson 1999; Tuovinen et al. 1980). Barium

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concentrations are lowest (mean value of 15 $\mu\text{g/L}$) in the drainage basins of the western Great Lakes and highest (mean value of 90 $\mu\text{g/L}$) in the southwestern drainage basins of the lower Mississippi Valley (EPA 2005c). Barium concentrations in the shallow aquifer below Denver, Colorado, have been reported to range from 18 to 594 $\mu\text{g/L}$ with a median value of 104 $\mu\text{g/L}$ (Bruce and McMahon 1996). Barium concentrations in most drinking water supplies are <200 $\mu\text{g/L}$ with a mean concentration of 28.6 $\mu\text{g/L}$ (EPA 2005c). In California, mean and median values of 302 and 160 $\mu\text{g/L}$, respectively, were measured for barium concentrations in drinking water supplies (Storm 1994).

Barium concentrations in groundwater supplies have been known to exceed EPA's maximum contaminant level (MCL) of 2.0 mg/L (2,000 $\mu\text{g/L}$) (EPA 2002a); this may be due to leaching and erosion of barium from sedimentary rocks (Calabrese 1977; Kojola et al. 1978). For example, community water supplies from deep rock and drift wells in northeastern Illinois have been found to have barium concentrations ranging from 1,100 to 10,000 $\mu\text{g/L}$ (Calabrese 1977). Many communities in Kentucky, Pennsylvania, and New Mexico have drinking water where the barium content is up to ten times higher than the MCL (EPA 2005c). Water samples taken from groundwater wells in Texas that are within 750 m of brine injection, dry, or plugged gas/oil wells contain barium ranging in concentration from 1.2 to 2,300 $\mu\text{g/L}$ (Hudak and Wachal 2001).

A mean concentration of 167 $\mu\text{g/L}$ for barium was measured in influent streams of a public waste water treatment plant in Melbourne, Australia (Wilkie et al. 1996). The amount of barium in the influent streams could not be accounted for based on the mean concentrations of barium in domestic water supplies (20 $\mu\text{g/L}$) or domestic sewage (38 $\mu\text{g/L}$). Instead, it is likely that the barium unaccounted for in the influent stream is the result of barium carried in effluents from industries that are discharged into the catchment area of the treatment plant.

Barium has also been found in sea water at concentrations ranging from 2 to 63 $\mu\text{g/L}$ with a mean concentration of 13 $\mu\text{g/L}$ (Bowen 1979).

Barium has been detected in surface water and groundwater samples collected at 257 and 561 of the 798 hazardous waste sites, respectively, where barium has been detected in some environmental medium (HazDat 2006). The HazDat information includes data from both NPL and other Superfund sites. Maximum concentrations of barium in surface water (lakes, streams, ponds, etc.) ranged from 0.33 to 18,100,000 $\mu\text{g/L}$ in 77 onsite samples (HazDat 2006). In comparison, maximum concentrations of barium in surface water (lakes, streams, ponds, etc.) ranged from 10 to 73,8000 $\mu\text{g/L}$ in 112 offsite

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samples (HazDat 2006). The maximum concentrations of barium in groundwater ranged from 0.064 to 2,100,000 $\mu\text{g/L}$ in 442 onsite samples (HazDat 2006). In comparison, maximum concentrations of barium in groundwater ranged from 0.05 to 803,000 $\mu\text{g/L}$ in 260 offsite samples (HazDat 2006).

6.4.3 Sediment and Soil

Barium is relatively abundant in the earth's crust and is found in most soils at concentrations (Table 6-3) ranging from about 15 to 3,500 ppm (dry weight) and mean values ranging between 265 and 835 ppm, depending on soil type (EPA 1995a; Kabata-Pendias and Pendias 1984; Lide 2005; Zenz et al. 1994). The barium content in cultivated and uncultivated soil samples collected during a number of field studies ranged from 15 to 1,000 ppm (mean concentration of 300 ppm) for B horizon soils (subsurface soils) in the eastern United States and from 70 to 5,000 ppm (mean concentration of 560 ppm) for B horizon soils in the western United States (Bowen 1979; Schroeder 1970; Shacklette and Boerngen 1984). Barium content ranged from 150 to 1,500 ppm for surface horizon soils collected in Colorado (mean concentration of 550 ppm) (Connor and Shacklette 1975). Soil samples (0–6 inch depth) taken from three New England cities, Boston, Providence and Springfield, were reported to have mean barium concentrations of 53.95, 45.29 and 45.17 mg/kg, respectively, and upper 95% interval values of 66.25, 59.43, and 51.03 mg/kg, respectively (Bradley et al. 1994). Soil samples were obtained from areas that were not influenced by industrial activity, such as along roads and sidewalks, parks and open lots, and may account for why the mean values for barium concentration were well below a mean value of 420 mg/kg for the United States.

Geometric mean concentrations of barium in sediments taken from 16 sampling sites along the southern shore of Lake Ontario and southeastern shore of Lake Erie ranged from 6.0 to 143.6 $\mu\text{g/g}$ (dry weight) (Lowe and Day 2002). Thirteen of the 16 sites had mean barium concentrations that exceeded EPA's guidelines (20–60 μg barium/g dry weight) for defining moderately polluted harbor sediments for this metal. However, these concentrations are lower than the mean barium concentration of 482.1 $\mu\text{g/g}$ in sediments collected from Lake Pontchartrain near New Orleans, Louisiana (USGS 2002c). The barium content in total suspended solids collected from the Mississippi River before it enters Lake Pontchartrain was 599 $\mu\text{g/g}$.

Barium concentrations in sediments near offshore drilling platforms are typically higher than unaffected sediments. Surficial and suspended sediments collected within 500 m of a drilling platform in the Santa Maria Basin offshore of central California contained barium at concentrations of 923 and 736 mg/kg dry

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Table 6-3. Concentrations of Barium in Surface Soils of the United States^{a,b}

Soil	Concentration ^c	
	Mean	Range
Sandy and lithosols on sandstone	400	20–1,500
Light loamy soils	555	70–1,000
Loess and soils on silt deposits	675	200–1,500
Clay and clay loamy soils	535	150–1,500
Alluvial soils	660	200–1,500
Soils over granite and gneisses	785	300–1,500
Soils over volcanic rocks	770	500–1,500
Soils over limestones and calcareous rocks	520	150–1,500
Soils on glacial till and drift	765	300–1,500
Light desert soils	835	300–2,000
Silty prairie soils	765	200–1,500
Chernozems and dark prairie soils	595	100–1,000
Organic light soils	265	10–700
Forest soils	505	150–2,000
Various soils	560	70–3,000
Mean concentration in Earth's crust ^d	500	—
Mean concentration in Earth's crust ^e	425	—

^aData obtained from Kabata-Pendias and Pendias (1984) unless indicated otherwise

^bData are for whole soil profiles

^cConcentrations expressed as ppm dry weight

^dZenz et al. 1994

^eLide 2000

Source: Adapted from EPA 1995a

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weight, respectively (Phillips et al. 1998). These values were higher (although not statistically significant) than the values of 869 and 687 mg/kg dry weight measured in surficial and suspended sediments, respectively, collected at a distance of 1,000 meters from the platform and are similar to the predrilling concentrations of barium in these sediments. In other California coastal sediments, for example the Southern California Bight, barium concentrations range from 145 to 1,259 ppm with an average of 720 ppm (Chow et al. 1978). Median barium concentration ranges in sediments from the lake system in Chiapas, Mexico were 54.4–121.2 and 50.3–155.3 µg/g dry weight in three lakes during the dry (June 2002) and rainy (September 2000) seasons, respectively (Pascual-Barrera et al. 2004). This lake system is an area of petroleum extraction and processing. Barium concentrations ranging from 180 to 2,800 µg/g dry weight (mean 729 µg/g dry weight) were reported in surface sediments (<63 µm fraction) collected in April 2002 from eight stations in Izmit Bay, Turkey (Pekey 2006).

Barium has been detected in soil and sediment samples collected at 369 and 260 of the 798 hazardous waste sites, respectively, where barium has been detected in some environmental medium (HazDat 2006). The HazDat information includes data from both NPL and other Superfund sites. Maximum concentrations of barium in soil (topsoil, <3 inches depth) ranged from 1.59 to 13,000 ppm in 84 onsite samples (HazDat 2006). In comparison, maximum concentrations of barium in soil (topsoil, <3 inches depth) ranged from 3 to 54,700 ppm in 28 offsite samples (HazDat 2006). Maximum concentrations of barium in sediment (lakes, streams, ponds, etc.) ranged from 13.1 to 17,600 ppm in 36 onsite samples (HazDat 2006). In comparison, maximum concentrations of barium in sediment (lakes, streams, ponds, etc.) ranged from 0.156 to 26,400 ppm in 92 offsite samples (HazDat 2006).

6.4.4 Other Environmental Media

Barium occurs in many foods at generally low levels (Table 6-4). In the Canadian Total Diet Study, the concentrations of barium were found to be less than 4 ppm (4,000 ng/g) in a variety of foods (Health Canada 2005). However, Brazil nuts have notably high concentrations of barium (3,000–4,000 ppm) (Beliles 1979). Some plants bioconcentrate barium from the soil (Beliles 1979; Reeves 1979; Schroeder 1970). The barium content in corn samples from Georgia, Missouri, and Wisconsin collected during a number of field studies ranged from 5 to 150 ppm with mean concentrations ranging from 15 to 54 ppm (Connor and Shacklette 1975). The barium content in other cultivated plants (e.g., lima beans, cabbage, soybeans, and tomatoes) from Georgia, Missouri, and Wisconsin ranged from 7 to 1,500 ppm with mean concentrations in various plants ranging between 38 and 450 ppm. The highest levels occurred in

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Table 6-4. Concentrations of Barium in Food Obtained from the Canadian Total Diet Study Between 1993 and 1999

Food Categories	Concentration (ng/g)	
	Average ^a	Range ^b
Milk	71.22	67.96–73.24
Dairy produce (ice cream, yogurt, cheese, cream)	332.37	70.61–962.93 ^c
Meats (beef, pork veal, lamb, organ meats)	131.82	12.06–237.57
Eggs	456.69	456.69
Poultry (chicken, turkey)	52.53	52.53
Fish (marine, fresh water, canned) and shellfish	137.28	36.17–481.34
Soups (meat, cream, tomato, dehydrated)	130.01	119.66–154.53
Breads, cereals, pasta, rice, pastries (cake, pies)	891.16	45.86–3,840.40 ^d
Vegetables	425.69	47.99–2,282.23 ^e
Fruits	570.33	57.62–3,750.03 ^f
Oils, fats, butter	32.45	20.67–53.08
Candy, syrups, jams, gelatin, puddings, honey, sugar	300.60	4.86–903.07 ^g
Peanut butter and peanuts	2,919.11	2,919.11
Beverages (beer, wine, coffee, tea, soft drinks, tap water)	70.94	13.05–151.82
Baby foods and formula	196.46	46.98–481.85
Frozen entrees	457.76	393.57–594.11
Processed foods (pizza, burgers, French fries, hot dog, etc.)	516.96	278.43–864.58

^aValues represent the average barium concentration in the foods covered under the individual food categories.

^bValues represent the range of average concentrations of the food items covered under the individual food categories.

^cThe highest barium concentrations were found in unprocessed cheeses (962.93 ng/g).

^dThe highest barium concentrations were found in wheat and bran cereals (3,840.40 ng/g), whole wheat bread (1,494.06 ng/g), muffins (1,434.30 ng/g), and cookies (1,029.13 ng/g).

^eThe highest barium concentrations were found in beets (2,282.23 ng/g) and carrots (1,309.25 ng/g).

^fThe highest barium concentrations were found in raspberries (3,750.03 ng/g) and strawberries (1,176.48 ng/g).

^gThe highest barium concentrations were found in chocolate candy bars (903.07 ng/g).

Source: Health Canada (2005)

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cabbage from Georgia and soybeans from Missouri and the lowest levels occurring in Georgia tomatoes (Connor and Shacklette 1975).

Grippo et al. (2006) measured various metal concentrations in dietary supplements purchased from random local vendors in the Little Rock, Arkansas area between 2002 and 2003. Barium concentrations in botanicals were 0.0200 and 15.4 ng/g in samples of milk thistle and kava kava, respectively. Barium concentrations in ephedra-containing supplements were 0.0400 and 93.3 ng/g in Virgin Earth and Xenadrine RFA-1, respectively. The authors noted that all metals measured in this study were detected at concentrations below toxic levels or physiological limits for daily intake, where such limits have been identified (Grippo et al. 2006).

The Wyoming Game and Fish Department collected game fish during the 2000–2001 season to survey the state's fisheries for metal contamination. Ninety-six fish composites (fillets) were collected, representing 11 species, from 28 lakes and reservoirs across Wyoming. In this study, barium concentrations were at or below the method detection limit of 0.05 mg/kg (Dailey et al. 2005). Mean barium concentrations ranging from 0.057 to 0.255 mg/kg wet weight were reported in muscle tissue of five species of sturgeons collected from the Caspian Sea (Pourang et al. 2005).

Barium is also found in anaerobic sewage sludge at concentrations ranging from 100 to 9,000 ppm (mean concentration of 800 ppm) and in aerobic sewage sludge at concentrations ranging from 100 to 300 ppm (mean concentration of 200 ppm) (Sommers 1977).

Barium concentrations in leachates from municipal landfills range from 0.11 to 9,220 µg/L (EPA 1990, 1991; Roy 1994).

Barium concentrations in fertilizers and soil amendments range from <0.2 to 669 µg/g mean (Raven and Loeppert 1997). The highest levels are in tilemsi phosphate rock (669 µg/g), austenite (408 µg/g), milorganite (165 µg/g), manure (153 µg/g), and compost (131 µg/g). There is some concern that continued use of fertilizers and soil amendments, which contain high amounts of barium and other metals, may result in an accumulation of barium in agricultural soils. The accumulation of barium in soils that is due to the continued use of fertilizers and soil amendments and the potential for increased content of barium in agricultural products and potential harm to the environment have not yet been assessed (Raven and Loeppert 1997).

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6.5 GENERAL POPULATION AND OCCUPATIONAL EXPOSURE

The primary routes of exposure of humans to barium are consumption of food and water and inhalation of ambient air (ICRP 1974; Reeves 1979; WHO 2001). Based on compliance monitoring data from the Federal Reporting Data System (FRDS), of the approximately 214 million people in the United States who are connected to a public water supply, it is estimated that about 150,000 people are exposed to barium concentrations greater than EPA's MCL of 2.0 mg/L (2,000 µg/L) (EPA 2002a). However, since 94% of all samples collected from public water supplies of the 100 largest cities in the United States had barium concentrations <100 µg/L (Durfor and Becker 1964), it is likely that most of the people connected to a public water supply receive drinking water with barium concentrations below the MCL. In a survey of drinking water from residences in EPA Region V (Indiana, Illinois, Michigan, Minnesota, Ohio, and Wisconsin) taken from the National Human Exposure Assessment Survey (NHEXAS) in 1995, an average barium concentration of 30 µg/L was obtained, with a 90th percentile value of 77 µg/L (Thomas et al. 1999). Assuming an average adult drinking water consumption rate of 2 L/day and that barium is present at concentrations of 30 µg/L, the average adult daily intake of barium through the consumption of drinking water would be 60 µg/day (0.86 µg/kg/day for a 70-kg adult). However, the International Commission for Radiation Protection (ICRP) estimates that the gastrointestinal absorption of barium is <5% (ICRP 1973).

The International Commission on Radiological Protection (ICRP 1974) has estimated that intake of barium through inhalation ranges from 0.09 to 26 µg/day. Based on reported urban air concentrations for barium (<0.005–1.5 µg/m³) (Tabor and Warren 1958) and assuming an average adult ventilation rate of 20 m³/day (EPA 1989), the calculated daily respiratory intake of barium ranges from <0.1 to 30 µg, which is comparable to the ICRP estimated intake range above. Based on the 8-hour time-weighted average threshold limit value (TLV) in workplace air of 500 µg/m³ (ACGIH 2004), and assuming an 8-hour inhalation of 10 m³ of air, a daily barium workplace intake of 5,000 µg can be calculated. NAS (1977) estimated that 75% of inhaled barium could be absorbed into the bloodstream if soluble barium salts were involved.

Since average ground level concentrations of an emission vary with the distance from the emission point, the population around a source site will be exposed to differing emission levels. Using an average population density of 27 persons/km² (based on actual population data from areas surrounding barium production and processing plants), it has been estimated that approximately 0–886 persons within an area of up to 32.8 km² around a source site could be exposed to soluble barium compound concentrations of

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>1.67 $\mu\text{g}/\text{m}^3$ in ambient air (Reznik and Toy 1978). Assuming that the average adult daily ventilation rate is 20 m^3 (EPA 1989), breathing these ambient air barium concentrations would result in daily respiratory intakes of >32 μg . No other correlations have been established between barium concentrations in air and geographical areas or land-use types.

The day-to-day intake of barium is likely to vary with the quantity and types of food ingested since the barium content in foods varies widely (Schroeder 1970). Based on consumption of food and beverages in long-term balance studies of four individuals, daily barium intake was estimated to range from 650 to 1,770 $\mu\text{g}/\text{day}$, or from 9.30 to 25.3 $\mu\text{g}/\text{kg}$ body weight/day based on an adult weight of 70 kg (Tipton et al. 1966, 1969). Assuming an estimated average barium intake of 60 $\mu\text{g}/\text{day}$ from drinking water that is based on the barium concentrations in drinking water obtained in the NHEXAS EPA Region V study (Thomas et al. 1999) and a consumption of 2 L of water per day, the barium intake from the consumption of non-drinking water dietary sources alone would range from 590 to 1,710 $\mu\text{g}/\text{day}$. Thus, food is typically the primary source of barium exposure for the general population. Gastrointestinal absorption of barium from food was reported to be approximately 6% (ranging from 1 to 15%) (ICRP 1974). However, reevaluation of this ICRP data and the data from other studies (Harrison et al. 1956; LeRoy et al. 1966); Tipton et al. 1969, Schroeder et al. 1972) using the methods of re-estimating barium absorption, which are based on current information of systemic kinetics of barium (Leggett 1992), suggest that gastrointestinal absorption of barium may be higher, generally ranging between 7 and 30% and could be as high as 95% in some individuals.

In the Canadian Total Diet Study (TDS) of 1993–1999, the average barium intake in individuals surveyed was found to be highest in young children (Health Canada 2005). The average barium intake ranged from 20.760 to 25.251 $\mu\text{g}/\text{kg}$ body weight/day for children ages 0–4 years old (Table 6-5). For individuals older than 4 years, the average barium intake decreased for both males and females with increasing age to values of 9.704 (20–39 years) and 7.839 (>65 years) $\mu\text{g}/\text{kg}$ body weight/day in males and 8.418 (20–39 years) and 7.546 (>65 years) $\mu\text{g}/\text{kg}$ body weight) in females. The average daily barium intakes from the Canadian TDS for males and females of all ages (8.817 $\mu\text{g}/\text{kg}$ body weight/day) is in reasonable agreement with the low end of the daily intake range for barium of 9.30 $\mu\text{g}/\text{kg}$ body weight/day determined by Tipton et al. (1966, 1969).

Mean daily balances (excluding loss via hair and sweat) determined from long-term balance studies of four adult subjects ranged from a negative balance of 800 μg to a positive balance of 890 μg (Tipton et al. 1966, 1969). Based on data from these studies, Schroeder (1970) estimated that human daily intake

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Table 6-5. Average Dietary Intake of Barium in Different Age/Sex Groups from the Canadian Total Diet Study (1993–1999)

Sex	Age	Intake ^a
Male and female	0–1 months	20.760
Male and female	2–3 months	23.350
Male and female	4–6 months	21.414
Male and female	7–9 months	21.213
Male and female	10–12 months	22.823
Male and female	1–4 years	25.251
Male and female	5–11 years	18.741
Male	12–19 years	11.759
Male	20–39 years	9.704
Male	40–64 years	8.976
Male	≥65 years	7.839
Female	12–19 years	9.280
Female	20–39 years	8.418
Female	40–64 years	7.855
Female	≥65 years	7.546
Male and female	All ages	8.817

^amicrograms barium per kilogram body weight per day

Source: Health Canada (2005)

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from food (1,160 μg), water (80 μg), and air (10 μg) would be approximately 1,250 μg , and that loss from urine (180 μg), feces (1,010 μg) and other sources (e.g., sweat and hair) (85 μg) would be 1,275 μg . Using these latter estimates of barium intake and loss, a negative barium balance of 25 μg would occur. According to ICRP, the average daily intake of barium from food and fluids (750 μg) and ambient air (0.09–26 μg) ranges from 750 to 776 μg . In addition, ICRP (1974) estimated that approximately 825 μg of barium is lost daily through the urine (50 μg), feces (690 μg), sweat (10 μg), and hair (75 μg). These intake and loss estimates indicate a negative daily balance of up to 75 μg . However, these negative daily balance values of 25 and 75 μg are not significant. Also, it is not expected that a negative daily balance would maintain a total body content of barium for a 70-kg adult of 22,000 μg , a value that was estimated from a study of barium content in major human organs and tissues (ICRP 1974; Schroeder et al. 1972). Ninety-three percent of this barium was found in bone and connective tissue. The remaining 7% of barium exists largely in fat, skin, and lungs.

Barium content in the human population has been determined in urine and major organs and tissues in more current studies. Barium concentrations in urine for the United States population aged 6 years and older were measured in the Third National Health and Nutrition Examination Survey (NHANES). The geometric mean (95% confidence interval) for the creatinine-adjusted levels of barium in urines for all ages was 1.44 (1.31–1.58) μg per gram of creatinine (CDC 2005). Within age groups, the geometric means for the barium concentration in urine decreased as a function of age, from 2.20 μg per gram of creatinine (6–11 years) to 1.45 μg per gram of creatinine (12–19 years) and 1.37 μg per gram of creatinine (20 years and older). The geometric mean concentration of barium in females (1.59 μg per gram of creatinine) was slightly higher than in males (1.30 μg per gram of creatinine). As a function of ethnicity, non-Hispanic whites had the highest geometric mean barium concentrations (1.62 μg per gram of creatinine) followed by Mexican Americans (1.18 μg per gram of creatinine) and non-Hispanic African Americans (0.891 μg per gram of creatinine). A median urinary concentration of 1,146 ng/L (range 295–5,250 ng/L) was reported in urine of 50 healthy individuals, aged 20–68 years, in central Italy (Alimonti et al. 2005).

Occupational exposure to barium primarily occurs in workers and miners who inhale barium sulfate (or the ore, barite) and barium carbonate dust during the mining of barite and the manufacturing and processing (e.g., mixing, grinding, and loading) of barium compounds (Beliles 1979; Reznik and Toy 1978; Schroeder 1970). Inhalation exposure to barium is also known to occur for industrial welders, especially those using barium-containing stick electrodes and self-shielded flux core wires, and those

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working in ceramic factories (Ramakrishna et al. 1996; Roig-Navarro et al. 1997; WHO 2001; Zschiesche et al. 1992).

Data from a workplace survey, the National Occupational Exposure Survey (NOES), conducted by NIOSH from 1980 to 1983, estimated the number of workers potentially exposed to various chemicals in the workplace in 1980 (NIOSH 1989a), including a separate tally of female workers. The data for barium and barium compounds included in the survey are summarized in Table 6-6. The NOES database does not contain information on the frequency, concentration, or duration of exposure of workers to any of the chemicals listed therein. This is a survey that provides only estimates of the number of workers potentially exposed to chemicals in the workplace.

6.6 EXPOSURES OF CHILDREN

This section focuses on exposures from conception to maturity at 18 years in humans. Differences from adults in susceptibility to hazardous substances are discussed in Section 3.7, Children's Susceptibility.

Children are not small adults. A child's exposure may differ from an adult's exposure in many ways. Children drink more fluids, eat more food, breathe more air per kilogram of body weight, and have a larger skin surface in proportion to their body volume. A child's diet often differs from that of adults. The developing human's source of nutrition changes with age: from placental nourishment to breast milk or formula to the diet of older children who eat more of certain types of foods than adults. A child's behavior and lifestyle also influence exposure. Children crawl on the floor, put things in their mouths, sometimes eat inappropriate things (such as dirt or paint chips), and spend more time outdoors. Children also are closer to the ground, and they do not use the judgment of adults to avoid hazards (NRC 1993).

The main exposures of children to barium are expected to occur mainly from the diet or by dermal contact with barium-containing dust, with minor exposures through barium in air. Data on the daily intake of barium in the total diet of children in the United States were not located in the available literature. However, the average daily intake of barium in children has been determined in a Canadian Total Diet Study (1993–1999), showing that children ages 0–48 months have the highest barium intake through their diet in comparison to older children (>4 years) and adults (Health Canada 2005). The average barium intake in young male and female children increased from 20.760 for infants (0–1 month) to 25.251 $\mu\text{g}/\text{kg}$ body weight/day for children ages 1–4 years (Table 6-5). For older children (>4 years), there is a continual decrease in the average daily barium intake, with values of 18.741 $\mu\text{g}/\text{kg}$ body weight/day for

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Table 6-6. Number of Workers Potentially Exposed to Barium and Barium Compounds

Chemical	Number of plants	Total workers (female workers)
Barium	815	10,308 (3,598)
Barium carbonate	4,494	61,019 (6,889)
Barium chloride	4,293	57,767 (15,249)
Barium hydroxide	1,423	35,351 (12,208)
Barium oxide (BaO ₂)	46	511 (325)
Barium nitrate	353	9,625 (2,699)
Barium sulfate	20,089	305,887 (83,800)
Barium sulfide	7	7 (0)
Chromic acid (H ₂ CrO ₄), barium salt (1:1)	20	3,546 (1,984)

Source: NIOSH 1989a

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individuals aged 5–11 years and then down to 11.759 $\mu\text{g}/\text{kg}$ body weight/day for males and 9.280 $\mu\text{g}/\text{kg}$ body weight/day for females aged 12–19 years. It is expected that the data obtained from the Canadian Total Diet Study will reasonably approximate the daily barium intake for children living in the United States. It is estimated that for children in the United States, the barium intake through drinking water will range between 36 and 60 $\mu\text{g}/\text{day}$. This estimate is based on an average concentration of 30 μg barium/L in drinking water within the United States (Thomas et al. 1999) and the consumption of 1.2–2.0 L water/day. A factor to be taken into account is that fractional intestinal absorption of metals in young children, as in young mammals, may be higher than in adults (Foulkes and Bergman 1993). Dermal contact with barium in household dust is not expected to result in uptake of barium through the skin. Oral intake of barium through hand-to-mouth exposures to barium-containing dust is likely to occur. However, it is not known how much barium is taken in through this route of exposure. There is also the potential of oral intake of barium through the licking or ingestion of crayons or water colors, but it is not known how much barium is ingested or how much is bioavailable (Rastogi and Pritzl 1996).

Dietary intake of barium in 3-month-old infants has been given by Biego et al. (1998) for exclusive consumption of various types of milk. The average barium intake from the consumption of breast milk only was determined to be 4 $\mu\text{g}/\text{day}$. Barium intake increases with exclusive consumption of bottled milk (39 $\mu\text{g}/\text{day}$), evaporated milk (42 $\mu\text{g}/\text{day}$), formula (44 $\mu\text{g}/\text{day}$), and dried milk (59 $\mu\text{g}/\text{day}$). The highest average intake of barium occurred with exclusive consumption of soya milk (91 $\mu\text{g}/\text{day}$). These intakes are based on an average daily intake of milk of 700 mL.

6.7 POPULATIONS WITH POTENTIALLY HIGH EXPOSURES

The general population is commonly exposed to barium primarily through ingestion of drinking water and consumption of food and beverages. However, certain populations face greater than average exposures to this element due to environmental sources, such as drinking water (EPA 1987). High levels of barium have been reported in groundwater from deep rock and drift wells in several communities in northeastern Illinois (Brenniman et al. 1981; Calabrese 1977) where barium is a naturally occurring geochemical pollutant found almost exclusively in the Cambrian-Ordovician Aquifer (Gilkeson et al. 1978). Other populations that might receive increased exposure to barium are consumers of crops grown on soils that have been used for the land farming of waste oil-well drilling muds (Bates 1988). Individuals who work at or live near barium mining, manufacturing, or processing plants might inhale higher ambient air concentrations or increased amounts of fugitive dust containing barium particulates. Populations living in the vicinity of the NPL sites known to be contaminated with barium may also be exposed to higher than

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background levels of the compound through contact with contaminated waste site media or barium in offsite air or water. Barium has been measured in air, surface water, and groundwater collected offsite of some NPL sites (HazDat 2006). No information was found regarding the sizes of these populations or their intake levels of barium.

6.8 ADEQUACY OF THE DATABASE

Section 104(i)(5) of CERCLA, as amended, directs the Administrator of ATSDR (in consultation with the Administrator of EPA and agencies and programs of the Public Health Service) to assess whether adequate information on the health effects of barium is available. Where adequate information is not available, ATSDR, in conjunction with NTP, is required to assure the initiation of a program of research designed to determine the health effects (and techniques for developing methods to determine such health effects) of barium.

The following categories of possible data needs have been identified by a joint team of scientists from ATSDR, NTP, and EPA. They are defined as substance-specific informational needs that if met would reduce the uncertainties of human health assessment. This definition should not be interpreted to mean that all data needs discussed in this section must be filled. In the future, the identified data needs will be evaluated and prioritized, and a substance-specific research agenda will be proposed.

6.8.1 Identification of Data Needs

Physical and Chemical Properties. The physical and chemical properties of metallic barium and its inorganic compounds have been well characterized (Boffito 2002; CHRIS Manual 2005; Dibello et al. 2003; DOT 2004; Genter 2001; HSDB 2007; Kresse et al. 2007; Lewis 1997; Lide 2005; NIOSH/OSHA 1978; NIOSH 1999; Budavari et al. 2001; OHM/TADS 1989; Parmeggiani 1983; Perry and Chilton 1973; RTECS 2007; Lewis 2000; Stokinger 1981; Weast 1989). Physical and chemical properties of organic compounds of barium have not been comprehensively examined probably due to the limited extent of formation of these compounds. However, further study of the properties of these compounds would help in understanding their role in the environmental fate and transport of barium, particularly at hazardous waste sites where high levels of organic contaminants might be present.

Production, Import/Export, Use, Release, and Disposal. According to the Emergency Planning and Community Right-to-Know Act of 1986, 42 U.S.C. Section 11023, industries are required to submit substance release and off-site transfer information to the EPA. The TRI, which contains this

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information for 2004, became available in May of 2006. This database is updated yearly and should provide a list of industrial production facilities and emissions.

Because barium compounds occur naturally and are widely used in oil well drilling muds, in steel, rubber and plastic products, glass and ceramics, chemical, and pyrotechnics industries, in insecticides, and as a smoke suppressant in diesel fuels (Bodek et al. 1988; Dibello et al. 2003; ILO 1983; Kirkpatrick 1985; Meister 2004; Stokinger 1981; Venugopal and Luckey 1978; WHO 2001; Worthing 1987), the potential for human exposure to these compounds, such as through ingestion of food and water or inhalation of ambient air, is substantial. Recent data on production volumes and import and export of barite and some barium compounds (e.g., barium chloride, barium carbonate, barium hydroxide, and barium oxide) are available (USGS 2006). In addition, only limited information on disposal of barium compounds was available (HSDB 2007; IPCS 1991; NIOSH/OSHA 1978). Additional information on production, import, export, and disposal would be useful in assessing the potential for the release of, and exposure to, barium compounds.

Environmental Fate. The partitioning of barium in environmental media is influenced by the specific form of the compound and such site-specific conditions as pH and cation exchange capacity (Bates 1988; Bodek et al. 1988; Bowen 1966; Giusti et al. 1993; Kabata-Pendias and Pendias 1984; Lagas et al. 1984; Tanizaki et al. 1992). Upon release to the environment, barium is most likely to partition to soils and sediments (Chow et al. 1978; DOE 1984; Rai et al. 1984; WHO 2001). Barium is transported in the atmosphere, surface waters, soil runoff, and groundwater. In surface waters and soils, barium may ionize and form various salts depending on the pH and the availability of anions (Bates 1988; Bodek et al. 1988; Bowen 1966; Kabata-Pendias and Pendias 1984; Lagas et al. 1984; WHO 2001). Additional information on the transport and transformation of barium in the atmosphere would be useful in developing a more complete understanding of the environmental fate of barium compounds.

Bioavailability from Environmental Media. Barium is absorbed following ingestion (Chou and Chin 1943; Cuddihy and Griffith 1972; McCauley and Washington 1983; Taylor et al. 1962) and inhalation (Cuddihy and Ozog 1973b). The bioavailability of barium from air, water, and food has been examined rather extensively in animals (Chou and Chin 1943; Cuddihy and Griffith 1972; McCauley and Washington 1983; Taylor et al. 1962) and humans (Tipton et al. 1969). However, bioavailability from soil has not been studied. Since soil is an important repository for barium, information on barium absorption from ingested soil would be useful in developing an understanding of the potential for exposure following ingestion of contaminated soils, particularly at hazardous waste sites.

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Food Chain Bioaccumulation. There is information that barium bioconcentrates in certain plants and aquatic organisms (Aruguete et al. 1998; Bowen 1966; Hope et al. 1996; IPCS 1991; Schroeder 1970; WHO 2001). However, the extent to which plants bioconcentrate barium from soil or to which uptake occurs in terrestrial animals is not well characterized. Further studies on the bioconcentration of barium by plants and terrestrial animals and on the biomagnification of barium in terrestrial and aquatic food chains would be useful to better characterize the environmental fate of barium and define the importance of food chain accumulation as a source of human exposure.

Exposure Levels in Environmental Media. Reliable monitoring data for the levels of barium in contaminated media at hazardous waste sites are needed so that the information obtained on levels of barium in the environment can be used in combination with the known body burden of barium to assess the potential risk of adverse health effects in populations living in the vicinity of hazardous waste sites. The need for additional information on the relationship between barium exposure and levels of barium achieved *in vivo* is essential if such concentrations are to be used as biomonitors of exposure.

Barium has been detected in the atmosphere (Bowen 1979; EPA 1984; Hildemann et al. 1991; IPCS 1991; Schauer et al. 1999; Shahin et al. 2000; WHO 2001; Winkler 2002), surface water (Barnett et al. 1969; Bowen 1979; DOI 1970; Durfor and Becker 1964; Durum and Haffty 1961; Elinder and Zenz 1994; EPA 2005c; Kopp 1969; Longerich et al. 1991; McCabe et al. 1970; Neal et al. 1996; Saleh and Wilson 1999; Tuovinen et al. 1980), groundwater (Bruce and McMahon 1996; Calabrese 1977; Hudak and Wachal 2001; Kojola et al. 1978), soils (Bowen 1979; Bradley et al. 1994; EPA 1995a; Kabata-Pendias and Pendias 1984; Lide 2005; Schroeder 1970; Shacklette and Boerngen 1984; Zenz et al. 1994), and foodstuffs (Beliles 1979; Connor and Shacklette 1975; Health Canada 2005; Schroeder 1970). There are reliable data to characterize the potential for human exposure via intake of drinking water (Durfor and Becker 1964; Hadjimarkos 1967; Thomas et al. 1999), and foods (Health Canada 2005; Tipton et al. 1966, 1969). Recent data on barium levels in plants and ambient air, soils, and groundwater, particularly from hazardous waste sites, would be useful in helping to develop a more complete understanding of the potential for human exposure.

Exposure Levels in Humans. Barium can be detected in blood, urine, feces, and biological tissues (CDC 2001, 2003; Mauras and Allain 1979; Schramel 1988; Shiraishi et al. 1987). However, there are no data correlating barium levels in tissues and fluids with exposure levels. Recent biomonitoring data exist for the U.S. general population (CDC 2005), although there are limited monitoring data for occupational

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exposure and for populations living near hazardous waste sites. This information is necessary for assessing the need to conduct health studies on these populations.

Exposures of Children. Child health data needs relating to susceptibility are discussed in Section 3.12.2, Identification of Data Needs: Children's Susceptibility.

Data on the exposure of children in the United States to barium are very limited. It is expected that the largest exposure to barium will be through the diet. Therefore, market basket surveys or total diet studies similar to those conducted by the U.S. Food and Drug Administration would be useful for providing data on typical levels of exposure via dietary intake for children in the United States. Data are available for barium intake in Canadian children obtained from a 1993–1999 total diet study (Health Canada 2005) and in a separate study (Biego et al. 1998) in infants (3 months old) from the exclusive consumption of breast milk and other types of milk.

Exposure Registries. No exposure registries for barium were located. This substance is not currently one of the compounds for which a sub-registry has been established in the National Exposure Registry. The substance will be considered in the future when chemical selection is made for sub-registries to be established. The information that is amassed in the National Exposure Registry facilitates the epidemiological research needed to assess adverse health outcomes that may be related to exposure to this substance.

6.8.2 Ongoing Studies

Three ongoing studies concerning the fate/transport of barium and measurement of barium in environmental media were identified in the Federal Research in Progress database (FEDRIP 2005). These studies are summarized in Table 6-7. No other pertinent ongoing studies were identified.

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Table 6-7. Ongoing Studies on Environmental Fate and the Potential for Human Exposure to Barium and Barium Compounds

Investigator	Affiliation	Research description	Sponsor
Batiza, R	Oregon State University	The proposed research will attempt to quantify the cold seep barite contribution to the marine sediment record by first quantifying the cold seep barite (an important carrier of barium) contribution to barium and radon fluxes within the San Clemente basin and then, secondly, quantify the effects of this source on the chemical signature of barite in the basin sediments.	NSF
Naehr, TM; MacDonald, IR	Texas A&M University Corpus Christi	Acquisition of a basic powder x-ray diffractometer system for qualitative and quantitative phase analysis in studies of (1) geologic materials to enhance research into the formation of diagenetic barite and silicate minerals in sediments from the Peruvian Continental Margin, (2) the authigenic seafloor deposits at sites of active submarine fluid expulsion in the Gulf of Mexico region to elucidate the geochemical environment at these sites, and (3) barium and other metals in soil and sediment samples.	NSF
Odom, JW	Auburn University	Develop analytical techniques for total and plant-available forms of barium and other metals in soils and for total analysis of these metals in plant material; determine the normal occurrence of both total and extractable forms of these elements in selected soil profiles; and ascertain the availability of soil test calibration data and soil test procedures for these elements.	Hatch

NSF = National Science Foundation

Source: FEDRIP 2005